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Critical phosphorus values from the Better Fertiliser Decisions for Pastures project: early insights from validation trials

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Abstract. Phosphatic fertilisers have made grazing in the south-west of Western Australia (WA) viable. However, there is evidence that a large proportion of pasture paddocks exceed soil test critical values at which 95% of maximum yield is achieved as identified in the national Better Fertiliser Decisions for Pasture (BFDP) project. Of 22 000 soil samples collected between 2009 and 2020, 56% exceeded the critical value for phosphorus (P), although there were constraints to potassium (K) and sulfur (S) and from soil acidity. Soils with available P exceeding the critical value are expected to lead to excessive losses of P to waterways, resulting in eutrophication. A trial program was established to validate the critical P values from BFDP so that concerns can be addressed about the relevance of these critical P values to WA conditions and to contemporary pasture varieties. Measured relative yields for 19 trials in the first year were mostly within 10% of that predicted from BFDP for soils with a P buffering index (PBI) >10. Soils with PBI <10 had measured relative yields up to 25% greater than predicted by BFDP, suggesting response calibrations for low PBI soils may require adjustment in the BFDP dataset. Some pasture yield gaps occurred when soil pH and P were low. Application of nitrogen (N), K and S almost doubled the yield when P was limiting or sufficient. Agronomic advice and practice should seek to optimise these multiple inputs, thereby optimising P use rather than applying P to levels above the critical value.

Keywords: Colwell P, critical values, pasture, phosphorus, phosphorus buffering index, soil testing.

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Introduction

Nutrient and fertiliser management on pasture soils is important from economic and environmental perspectives. Fertilisers can account for a significant proportion of farm input costs, and the egress of nutrients from agriculture and impacts on riverine and estuarine water quality are of increasing public concern (Melland *et al.* 2008; Gourley and Weaver 2012). These factors highlight the importance of using an evidence-based approach to fertiliser decision making, contingent on soil test critical values derived from

nutrient response trials. An evidence-based approach will lead to improvement of on-farm nutrient use efficiency, minimised offsite nutrient loss risk, and optimised return on fertiliser investment.

The projects Better Fertiliser Decisions for Pastures (BFDP; Gourley *et al.* 2019) and Better Fertiliser Decisions for Crops (BFDC; Speirs *et al.* (2013) focused on collation, filtering and packaging of existing trial data rather than collection of new data. In both BFDP and BFDC, significant historical trial data were contributed from

government agencies and fertiliser companies around Australia. For example, the BFDP database consists of data from 3000 experimental years, 250 experiments, 1600 field sites, and 48 000 pasture yield measures. BFDC has been developed into a web-based system that allows authorised users to supplement the dataset with more results as they become available, whereas BFDP does not have this facility. This may contribute to a perception of a lack of currency of the BFDP data. The outcome of BFDP was a suite of critical soil test values for phosphorus (P), potassium (K) and sulfur (S) for grazed pastures in Australia.

Nutrient and fertiliser recommendations for crops and pastures are based on the predictive functions and critical soil test values of decision support systems (DSS) derived from nutrient response trials. Some landholders and industry stakeholders are reluctant to base fertiliser decisions on the evidence provided in studies such as BFDP. This is despite many years of trials, meta-analysis of nationally aggregated trial information (Speirs *et al.* 2013; Gourley *et al.* 2019), peer-reviewed scientific publications, and the utilisation of critical values in DSS and extension programs. Instead, there is a preference to adhere to traditional fertiliser practice of 'one bag of superphosphate per acre per year' (i.e. 200 kg superphosphate/ha.year) because of strong memories of responses of pasture to P applications on infertile agricultural land in south-west Western Australia (WA) in the 1960s. Other factors cited for this reluctance also include the lack of contemporary pasture varieties used in the trials that contributed to BFDP, with concerns that the newer cultivars may have higher P requirements than earlier cultivars, and that the trials were conducted from 1955 to 2006 and are therefore not contemporary from a scientific and extension perspective, and therefore the derived critical values are not relevant to conditions in south-west WA.

When assessed against critical values determined from studies such as BFDP, records of soil tests reinforce the notion that there is a tendency for landholders to follow traditional fertiliser practice rather than use an evidence-based approach (Weaver and Reed 1998; Weaver and Wong 2011; Gourley *et al.* 2019). For example, Weaver and Reed (1998) showed that of 7950 soils sampled (0–10 cm) on the south coast of WA in 1988 and 1989, 49% had high P status when assessed against critical Colwell P (Colwell 1965) values determined from pasture trials (Yeates 1993). Weaver and Wong (2011) demonstrated that of 109 000 soil sample records (0–10 cm) collected in south-west WA during 2008–10, 57–69% exceeded the critical Colwell P value to achieve 90% of relative yield (RY) for pastures, and 39–93% exceeded the critical Colwell P value to achieve 90% of RY for crops, depending on soil P buffering index (PBI; Burkitt *et al.* (2002). Weaver and Wong (2011) also showed that of 2160 soil samples (0–10 cm) from Australian dairy farms collected in 2007 and 2008, 80–95% exceeded the critical Colwell P value to achieve 95% of RY for pastures.

These findings from an assessment of soil test records belie the notion that an evidence-based approach to fertiliser decision making is ubiquitous. It is clear that approaches beyond the collection and publication of data from nutrient

response trials are required to persuade landholder and industry stakeholders that production or cost effectiveness of fertiliser use does not increase when soils are fertilised to levels above critical values (Simpson *et al.* 2009; Gourley *et al.* 2019). With this background, a project called 'uPtake', supported by a technical reference group with representatives from government departments, catchment groups, grazing and fertiliser industry groups, research scientists and farmers, is conducting pasture trials to derive PBI-specific soil test P response calibrations and critical values for pastures grown at >600 mm rainfall in south-west WA. Results from 19 trials conducted in the first year are presented and compared with data from BFDP (Gourley *et al.* 2019). The primary objective of the trials was to validate the critical P values in BFDP. Secondary objectives were to deliver behavioural change outcomes including: (i) landholders and industry stakeholders accepting and using critical soil test P values to support evidence-based fertiliser recommendations; (ii) increased profitability for landholders; and (iii) improved water quality by reducing excessive P levels in the soil. A scientific framework to allow comparison with BFDP was central to achieving the primary objective, and to add credibility to the secondary objectives.

Materials and methods

Background

The present study has limited resources, time and trials for achieving outcomes similar to and outputs as extensive as BFDP. Hence, the current understanding of the science behind the responsiveness of soils with different PBI and levels of soil P fertility (Gourley *et al.* 2019) was applied to the selection of trial locations so that P response calibrations could be developed and comparisons with BFDP drawn. Gourley *et al.* (2019) notes the importance of soil PBI to the determination of critical Colwell P values for pastures. A similar meta-analysis for crop nutrient response relationships in Australia could find no PBI-dependent critical P values (Bell *et al.* 2013), which the authors attributed to insufficient data and a range of knowledge gaps relating to current cropping practices, including minimum tillage and soil characterisation. In addition, BFDC lists the minimum requirements for a trial to qualify for inclusion in its national dataset. Some essential requirements include records of: (i) site location; (ii) crop type; (iii) experimental design; (iv) soil sampling depth; (v) soil test method and the units reported; (vi) mean yield (t/ha) for each treatment; (vii) Y_0 , Y_{\max} and the equation fitted treatment yields, where Y_0 is the mean yield from the control and Y_{\max} is the maximum yield from a fitted response equation or from the maximum nutrient rate depending on the trial design.

Site selection framework

Soil test data from an annual soil-sampling program guided the selection of potential trial sites. A framework (Fig. 1) based on PBI and P fertility groups enabled the selection of sites for development of nutrient response calibrations for each PBI group. Historical soil-sampling programs provided

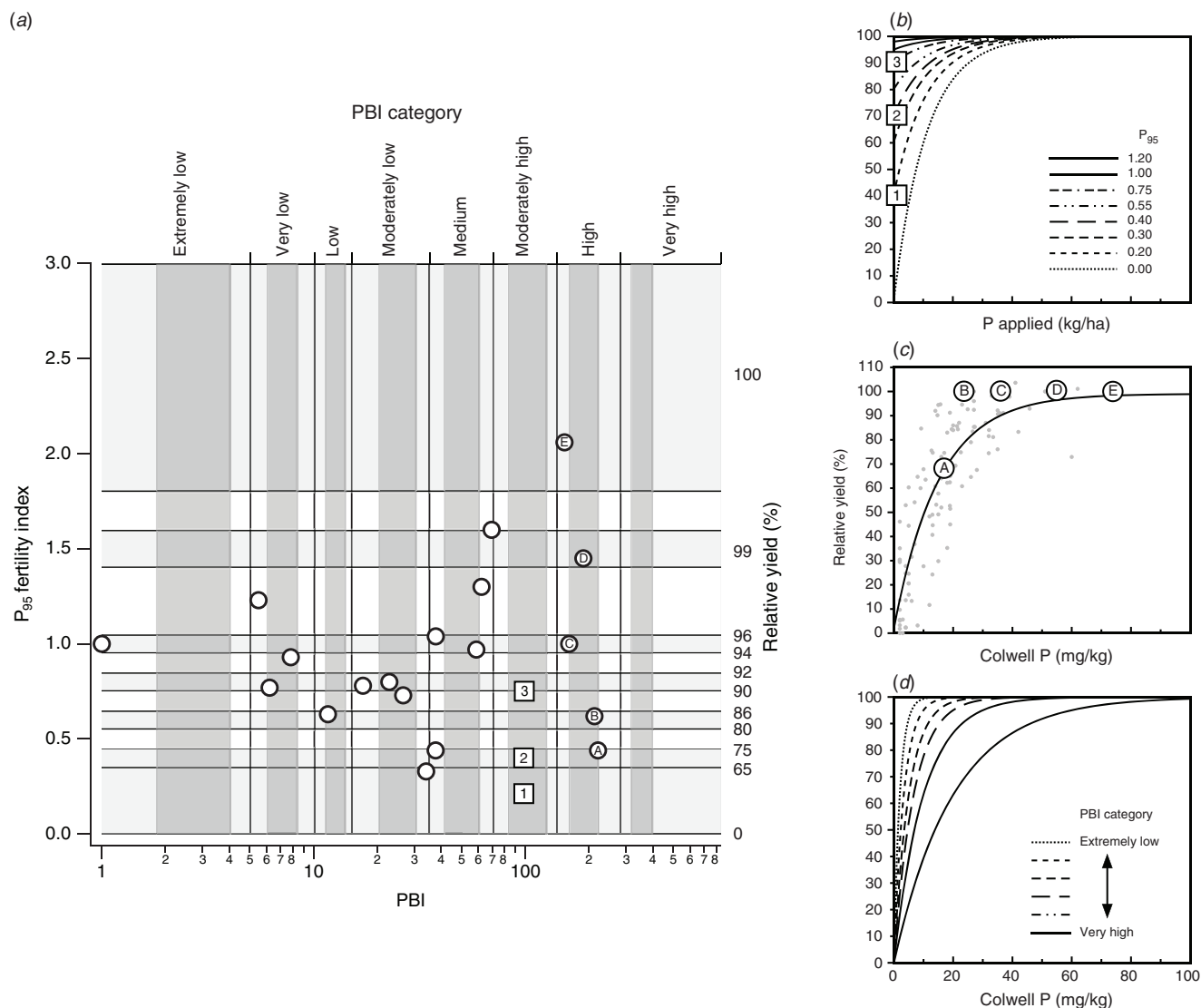


Fig. 1. (a) Framework for selecting trial sites, with consideration of soil PBI and soil P_{95} fertility index. Framework shows PBI categories (Gourley *et al.* 2019), further narrowed to limit category edge effects (dark grey vertical bands). Horizontal differentiation by targeted P_{95} fertility index or RY ranges (light grey bands). Intersection between dark grey and light grey bands identifies targeted characteristics of trial sites for inclusion in the program. Circles show the characteristics of trial sites used in 2019. (See part b for reference to numbered squares, and part c for reference to lettered circles.) (b) Conceptual P response curves (based on curvature values from Gourley *et al.* 2019) showing the expected P response within the 'moderately high' PBI category for trials with different initial P_{95} fertility index values. Numbered squares show the expected relative yield from control plots with these initial P_{95} fertility index values. (c) Single P response calibration aggregating trials for Colwell P and RY from within the 'high' PBI category (lettered circles A–E) overlaying BFD data (●) for the 'high' PBI category with fitted response calibration for BFD data. (d) Conceptual P response calibrations for soils within different PBI categories (Gourley *et al.* 2019), showing the expected effect of soil PBI on the curvature of P response calibrations. Trials conducted on soils with different PBI and with P_{95} fertility index of zero would be expected to show similar differences in curvature.

information on the likelihood that sites with the required PBI and P fertility characteristics could be identified (Table 1). These data were skewed towards high P fertility and towards medium–high PBI, with 56% of soils showing a P_{95} fertility index >1 (see below for explanation of index). The lowest and highest PBI groups were not well represented (Table 1). The 19 field sites chosen for the first year of the 4-year program partially represented the range of P fertility and PBI (Fig. 1a) required to develop response calibrations.

Prior to trial establishment, soil samples (0–10 cm) were collected to determine initial site conditions such as PBI, Colwell P and K (Colwell 1965), pH (in CaCl_2 , pH_{Ca}), and KCl-40S (Blair *et al.* 1991). A soil P fertility index was calculated (Cope and Rouse 1973; Simpson *et al.* 2011). This index simplifies soil P fertility interpretation and removes the need to use a complex array of PBI-specific critical Colwell P values which increase with increasing soil PBI (Gourley *et al.* 2019). The index was calculated as the

Table 1. Percentage of soil samples within specified PBI and P₉₅ fertility index ranges taken from 22 000 soil samples collected from the coastal plain of south-west WA 2009–20Intensity of shading indicates degree of representation; green shading for individual combinations PBI or P₉₅ fertility index groups, and red shading within PBI or P₉₅ fertility index groups

PBI	PBI category	P ₉₅ fertility index							Total
		0.0–0.25	0.25–0.50	0.50–0.75	0.75–1.0	1.0–1.25	1.25–1.50	>1.50	
0–5	Extremely low	0.02	0.17	0.29	0.46	0.29	0.33	0.76	2.31
5–10	Very low	0.05	0.34	0.43	0.75	0.26	0.37	0.42	2.62
10–15	Low	0.07	0.50	0.52	0.90	0.34	0.38	0.39	3.09
15–35	Moderately low	0.18	2.10	2.63	3.68	1.76	1.92	2.11	14.4
35–70	Medium	0.11	1.50	2.33	4.12	2.47	2.84	6.77	20.1
70–140	Moderately high	0.17	1.77	2.33	4.06	2.24	2.70	9.34	22.6
140–280	High	0.09	1.85	2.71	4.23	2.47	3.33	8.94	23.6
280–840	Very high	0.10	1.18	1.60	2.55	1.26	1.42	2.98	11.1
>840	Extremely high	0.01	0.06	0.02	0.01	0.00	0.01	0.01	0.127
Total		0.81	9.48	12.9	20.8	11.1	13.3	31.7	

ratio of pre-trial Colwell P to critical Colwell P (Gourley *et al.* 2019) to achieve 95% RY and is referred to here as P₉₅ fertility index. Soils with a P₉₅ fertility index of 1 should achieve 95% RY assuming no other constraints, whereas soils with a P₉₅ fertility index <1 should show response to applications of P depending on the degree of deficiency, and those with a P₉₅ fertility index >1 should be non-responsive to P applications. The P₉₅ fertility index (Eqn 1) can be applied at any target RY, but in most studies a target of 90% or 95% RY is used:

$$P_{\text{target RY}} = \frac{\text{Measured Colwell P}}{\text{Critical Colwell P for target RY}} \quad (1)$$

Within the framework, narrowed ranges of P₉₅ fertility index and PBI were defined (Fig. 1) to limit the likelihood that analytical variability could place a trial site in a different PBI group at the boundaries of each PBI range. Equally, the P fertility ranges recommended in the framework span the region where a P response calibration is likely to be well defined, and are an attempt to follow the BFDC criteria that at least three points must fall in the 80–95% RY range.

At each site, a P response curve was developed, with a higher intercept expected for sites with higher starting P fertility (Fig. 1b). Within each PBI category, the relative response to P was compared with the BFDP relationship and data (Fig. 1c). Each trial contributes a single point consisting of an initial Colwell P and PBI value from soil testing, and a RY measurement from the trial to the response calibration. The current dataset represents the first of 4 years of trials. After 4 years, the response calibrations will be compared with the series of BFDP curves where the curvature varies with PBI (Fig. 1d).

Trial treatments and management

Sites were seeded with pasture varieties in consultation with the individual farmers about which varieties they wished to

plant that were appropriate for their soil type and rainfall zone. Pastures were sown with grower equipment in mixes that included ryegrass and legume clover. Where necessary, the existing pasture was supplemented with pasture varieties that included mixes of oats; Winterhawk, Vortex, Ascend and Arnie annual ryegrass; Gosse, Narrikup, Dalkeith and Denmark subterranean clover; Lightning Persian clover; Zulu II arrowleaf clover; and Vista balansa clover. Knockdown herbicides were used before sowing, and selective herbicides were applied if required during the season to control weeds. Sites were sown from mid-April through to mid-June as rainfall permitted. Stock exclusion was achieved with fencing, the timing of which was dependent on trial type (Table 2). Trial design 1a was fenced as soon as it was established. Trial designs 1b and 2 allowed stock access to the trial until late August or early September (depending on pasture development), when fencing was erected and stock were excluded (Table 2).

Trial designs 1a and 1b were a randomised block design with three replicates of each treatment, whereas trial design 2 was randomised split plot (Table 2). Designs 1a and 1b included five rates of P from 0 to 40 kg/ha with basal nutrients (nitrogen (N), K, S, copper (Cu), zinc (Zn)) and two rates of P (0 and 40 kg/ha) without the basal nutrients. Plots were 2.2 m by 20 m with a buffer of 0.55 m between treatments. There were 12 trials with design 1a or 1b: seven trials were fenced for the entire season and mown every 5–6 weeks for biomass determination (1a), and five trials were grazed until spring then locked up for biomass measurements (1b). Seven trials used design 2 and included three rates of P (0, 5 and 40 kg/ha); the plots were split and basal nutrients (N, K, S, Cu, Zn) were applied to one half. Design 2 plots were narrower than design 1a and 1b plots, and were managed by allowing stock access until spring lockup before biomass measurements.

All P treatments were achieved with the application of All Phos fertiliser (CSBP), selected for its low S content (20.3% P,

Table 2. Trial designs, treatments and management

Design	P applied (kg/ha)	Basal nutrients at establishment (kg/ha)	Basal nutrients in season (kg/ha)	Management	No. of trials	Plot dimensions (m)	No. of plots
1a	0, 5, 10, 20, 40	N (59.8), K (49.9), S (22.1), Cu (3.5), Zn (2.4)	N (60), K (32.4), S (22.9); after each cut (approx. each 6 weeks)	Stock excluded	7	2.2 by 20	21
1b	0, 40	No basal	No basal	Stock access until spring lockup	5	2.2 by 20	21
	0, 5, 10, 20, 40	N (59.8), K (49.9), S (22.1), Cu (3.5), Zn (2.4)	N (60), K (32.4), S (22.9); at lockup and late September				
	0, 40	No basal	No basal				
2	0, 5, 40	N (59.8), K (49.9), S (22.1), Cu (3.5), Zn (2.4)	N (60), K (32.4), S (22.9); at lockup and late September	Stock access until spring lockup	7	2 by 20	18
	0, 5, 40	No basal	No basal				

1.0% S). Design 1 trials had a basal fertiliser (N, K, S, Cu, Zn) applied at establishment, and pasture was cut to 5 cm with a mower at intervals of ~6 weeks from June to September, depending on growth rates. After every cut, in-season applications of N, K and S basal fertiliser were applied (Table 2). Trials using designs 1b and 2 had a basal fertiliser (N, K, S, Cu, Zn) applied at establishment of trial. Two further in-season applications of N, K and S basal fertiliser were applied, one at lockup and one in late September (Table 2).

For design 1a trials, the entire site was mown following each biomass assessment and the pasture removed to simulate grazing, whereas trials using designs 1b and 2 retained stock access until spring, when the site was locked up to enable a flush of growth during spring for subsequent biomass assessment.

Pasture measurements

Pasture growth was determined by weighing strips of mown pasture of known length and width within the plots. Wet pasture was weighed, the weight recorded, and a subsample of known weight retained and dried to determine dry matter.

Percentages of clover, grass and weeds were assessed using the BOTANAL method as described by Cayley and Bird (1996). Sites of type 1a design were assessed twice in the season (once at the first biomass assessment and then again in October). Trials with designs 1b and 2 were assessed once only, in October.

Species-specific tissue testing was conducted for all sites at or just before 10% flowering (mid-late September, depending on site growth). If composition of the pasture sward was >30% legume, the legume component was sampled. If the legume composition of the sward was <30%, ryegrass was sampled. Species-specific tissue testing was chosen owing to budget constraints, and although plant tissue results are of value in assisting with interpretation, they were secondary to the primary objective of validating critical P values from BFDP. Plant tissue samples were analysed for P, K, S, calcium, magnesium, sodium, iron and boron, using the methods described by McQuaker *et al.* (1979). In brief, a mixture of hydrogen peroxide and nitric acid is added to a dry plant sample and heated until completely digested. Digests are

then read by inductively coupled plasma spectroscopy, which determines total elements present within the plant. Plant tissue was analysed for N by method 9G2 of Rayment and Lyons (2011). In this method, plant samples are analysed for total N via the Dumas high-temperature combustion method in a LECO analyser. Samples are loaded into a combustion tube at 950°C and flushed with oxygen. Gases generated from this process are measured for N by using a thermal conductivity cell.

Data analyses

Several non-linear statistical relationships have been used to describe yield response. The most successful of these is the Mitscherlich equation (Eqn 2), which was applied to the trial data when there were five rates of P in the trial design (Table 2):

$$Y = A \times (1 - B \times \exp(-C \times X)) \quad (2)$$

where Y is plant yield, absolute (t/ha) or relative (%); A is Y_{\max} (nutrient non-limiting); B is site responsiveness, $(A - Y_0)/A$, where Y_0 is yield when $X = 0$ and ranges from 0 to 1; C is curvature coefficient; and X is amount of nutrient measured in the soil test (mg/kg) or applied (kg/ha).

When there were at least three P rates in the trial design (Table 2), RY was estimated using Y_0 (yield from the control) and Y_{\max} (yield from maximum P rate).

Analysis of variance (ANOVA) including interactions was undertaken with GENSTAT 20th Edn (VSN International) to detect differences ($P < 0.05$) in dry matter produced for the treatments for each trial. Blocking was included as a factor for the trials with randomised block designs (1a and 1b) to account for spatial variation, but not for trials with design 2 because of trial layout constraints. A post hoc analysis determined least significant differences (l.s.d.) (Gramm *et al.* 2007).

Relative yield was then determined from Eqn 3:

$$RY = \frac{Y_0}{Y_{\max}} \times 100 \quad (3)$$

where Y_0 is pasture yield with no nutrient applied, and Y_{\max} is maximum pasture yield when non-limiting nutrient is applied.

Collated Colwell P, PBI, P₉₅, RY and ANOVA information was used to classify the trials as responsive or non-responsive,

and RY was compared with the predicted RY from BFDP (eqns 6 and 7 from Gourley *et al.* 2019; referred to here as Eqns 4 and 5):

$$RY = 100 - 100 \times \exp \left((-0.196 + 0.046 \times PBI^{0.179} \times \text{Colwell P}) \right) \tag{4}$$

$$RY = 100 - 100 \times \exp \left((-0.196 + (0.045 - 0.227 \times \exp(-0.201 \times PBI) \times PBI^{0.179} \times \text{Colwell P}) \right) \tag{5}$$

Equation 4 modified the Mitscherlich equation (Eqn 2) to include a *c* coefficient that varied according to PBI, and Eqn 5 made further adjustments to the *c* coefficient to account for other published work suggesting lower critical Colwell P values and greater responsiveness for sites that had low PBI (Yeates 1993; Moody 2007). A trial was deemed non-responsive if ANOVA determined no significant response to P application. In these cases where a Mitscherlich fit was used, RY was assigned 100% for trials that showed no response to P application.

The trial results (Colwell P and RY pairs) were also compared with the BFDP data by plotting 95% prediction intervals (Helsel and Hirsch 1992) associated with Colwell P and RY pairs from the BFDP data for PBI ranges in which the trial results were situated (Fig. 1a, c). This allows an assessment of whether a new observation (i.e. these trials) is likely to have come from the same distribution as previous data (BFDP) or from a different distribution.

Estimates of P removal in harvested biomass were made by multiplying median dry matter yield by the median P concentration of sampled plant tissue for trial treatments.

For trial designs 1b and 2, which maintained stock access until spring lockup, it was assumed that only 20% of the P in harvested biomass was removed in stock (Weaver and Wong 2011), whereas for trial design 1, 100% of the P was assumed to have been removed in harvested dry matter.

Results

The PBI of the selected sites corresponded with the PBI categories most represented by soil types in the study area (Table 1): moderately low, medium and high PBI categories. The other PBI categories are yet to be well represented in this study. The medium and high PBI categories included a wide range of P₉₅ fertility index soils, whereas the extremely low PBI category included only one site and both the moderately high and very high PBI categories none (Fig. 1).

Table 3 presents initial soil characteristics and whether a P response was predicted at each trial site. Response of mean annual cumulative dry matter to P applied is shown for typical trials with a P response (Fig. 2a) and without a response (Fig. 2b). The responsive trial (Fig. 2a) showed a significant (*P* < 0.05) increase in dry matter at rates of applied P from 5 to 40 kg/ha. The RY of the control (0 kg P/ha) was almost identical with basal nutrients applied (47.8%), without basal nutrients applied (50.8%), and determined from a Mitscherlich curve fit (51.1%). The dry matter yield increase from the addition of basal nutrients was significant (*P* < 0.05), and similar for applied P rates of 0 kg P/ha (76%) and 40 kg P/ha (87%) (Fig. 2a).

The median P concentration of subterranean clover sampled from the typical responsive trial at 0 kg P/ha with and without basal nutrients was 0.13% and 0.14%, respectively, resulting in P removal in dry matter of 4.04 and 2.41 kg/ha. For the same

Table 3. Initial soil conditions for each trial along with P₉₅ fertility index and prediction of a P response
Predicted response was estimated using Eqn 8 from Gourley *et al.* (2019) to determine a critical Colwell P for 95% RY at the measured PBI, and Eqn 1 (this paper) to determine the P₉₅ fertility index. A threshold P₉₅ fertility index of 1 determined whether or not the site was predicted to respond to P. This approach recognises the difficulty in measuring a statistically significant response beyond a RY of 95%

Trial	Trial design	Texture	PBI	Colwell P	Colwell K (mg/kg)	KCl-40 S	pH (in CaCl ₂)	Measured P ₉₅ fertility index	Predicted response
1	1a	Sand	1	9	79	28.4	4.2	1.00	N
2	1a	Sand	6.2	10	48	6.3	4.7	0.77	Y
3	2	Sandy loam	7.8	17	144	7.5	4.7	0.93	Y
4	1b	Sand	5.5	16	47	115.2	5.0	1.23	N
5	2	Sandy loam	11.7	14	175	7.3	5.4	0.63	Y
6	1b	Sand	17.1	18	31	91	4.7	0.78	Y
7	2	Sandy loam	22.8	20	208	8.3	4.5	0.80	Y
8	1a	Sandy loam	34	9	50	15.6	4.9	0.33	Y
9	2	Sandy loam	26.5	19	177	7.7	4.7	0.73	Y
10	2	Sand	37.8	28	93	7.3	5.7	1.04	N
11	1a	Sandy loam	37.7	12	88	5.2	4.1	0.44	Y
12	2	Sandy loam	58.7	28	63	8.5	4.7	0.97	Y
13	1b	Sandy clay loam	62.1	39	56	7.8	6.2	1.30	N
14	1b	Sandy clay loam	69.3	48	155	13.3	5.4	1.60	N
15	2	Sandy loam	161.2	36	137	10.4	5.4	1.00	N
16	1a	Sandy clay loam	220.6	17	137	18	4.4	0.44	Y
17	1a	Sandy clay loam	211.9	24	163	86.8	5.8	0.62	Y
18	1b	Sandy clay loam	188	55	99	14.6	5.1	1.45	N
19	1a	Sandy clay loam	153.2	74	42	12.4	4.7	2.06	N

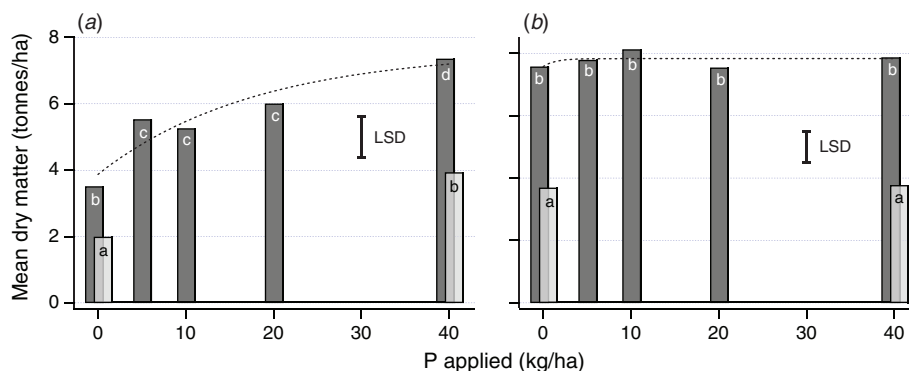


Fig. 2. Cumulative seasonal dry matter response to P application, with basal nutrients (dark grey bars) and without basal nutrients (light grey bars) applied for a typical (a) responsive site (trial 11) and (b) non-responsive site (trial 19). Dashed line shows Mitscherlich curve fit through treatments with basal nutrients applied. Means with the same letter are not significantly different ($P > 0.05$); l.s.d. ($P = 0.05$) range shown on each graph.

trial when 40 kg P/ha was applied, the median P concentration with and without basal nutrients was 0.23% and 0.26%, respectively, resulting in P removal in dry matter of 17.29 and 10.34 kg P/ha.

The median P concentration of ryegrass sampled from the typical non-responsive trial at 0 kg P/ha with and without basal nutrients was 0.26% and 0.31%, respectively, resulting in P removal in dry matter of 19.95 and 11.19 kg P/ha. For the same trial when 40 kg P/ha was applied, the median P concentration with and without basal nutrients was 0.32% and 0.38%, respectively, resulting in P removal in dry matter of 26.22 and 13.33 kg P/ha.

The non-responsive trial showed no significant ($P > 0.05$) increase in dry matter at any P application rate (Fig. 2b). The RY of the control (0 kg P/ha) was almost identical with basal nutrients applied (96.2%), without basal nutrients applied (97.7%), and determined from a Mitscherlich curve fit (100%). The dry matter yield increase from the addition of basal nutrients was significant ($P < 0.05$), and similar for applied P rates of 0 kg P/ha (104%) and 40 kg P/ha (108%) (Fig. 2b).

Of the 19 trial sites, 14 sites did not respond to P applications, and five sites did. Of the 11 trial sites predicted to respond, five sites were responsive to P application (e.g. Fig. 2a), and six sites were not (e.g. Fig. 2b). All five trials that did respond were predicted to respond based on the models in BFDP (Gourley *et al.* 2019). Of the eight trial sites predicted not to respond to P, none responded. Fifteen trials were responsive to the addition of basal nutrients with or without the addition of P, and 12 trials were responsive to basal nutrients alone without the addition of P. None of the trials showed a significant interaction between basal nutrients and P rate.

For the same P supply, whether that be from the soil alone (0 kg P/ha), or from soil with applied P at 40 kg/ha, the addition of basal nutrients increased dry matter and P removal in biomass, and therefore P demand by >100% (Fig. 2a, b). Omitting basal nutrients caused a yield gap of almost 50% in both 0 kg P/ha and 40 kg P/ha treatments (Fig. 2a, b). This

significant effect of basal nutrients was seen to varying degrees in 15 of the 19 trials. The dry matter increase ranged from 141% to 267% for 0 kg P/ha, and from 137% to 244% for 40 kg P/ha.

Most sites did not respond to applications of P even when they were predicted to be responsive on the basis of soil test results (Fig. 3a, Table 3). The $P_{0.5}$ fertility index ranged from 0.33 to 0.73 at responsive sites, and from 0.62 to 2.06 at non-responsive sites. The RY measured in the trials, compared with Eqn 5, was within 10% of the predicted response based on soil test results in most cases (Table 3, Fig. 3b). Non-responsive trials always had a positive RY difference, whereas responsive trials always showed a negative RY difference. Non-responsive trials had RY differences of up to +15%. Similar patterns were observed when RY differences were estimated by using Eqn 4; however, as PBI decreased to <15, the RY differences increased by as much as 20%.

Three responsive trials had a difference in RY of -20%, and at two of these sites, pH_{Ca} was 4.7 (site 9) and 4.1 (site 11) (Table 3). The mean pH of responsive sites was 4.7, and non-responsive sites 5.1, but this difference when the data was grouped was not significant ($P > 0.05$).

The BFDP data and 95% prediction intervals for the same overlapping PBI ranges specified by Gourley *et al.* (2019) are shown in Fig. 4 with Colwell P and RY pairs from this study. All 19 trials fall within the 95% prediction interval boundaries from BFDP.

Discussion

Results from the 19 trials to date are consistent with the national BFDP data and within the 95% prediction interval bands (Fig. 4), suggesting that the trials are in the same population (Helsel and Hirsch 1992). The trials are consistent with the expected P response (Fig. 3), and critical soil test values derived using Eqn 5. Although only five of 11 trials expected to respond to P showed a P response, the range of 95% prediction intervals (Fig. 4) from a large number of trials suggests that it would not be unusual in a

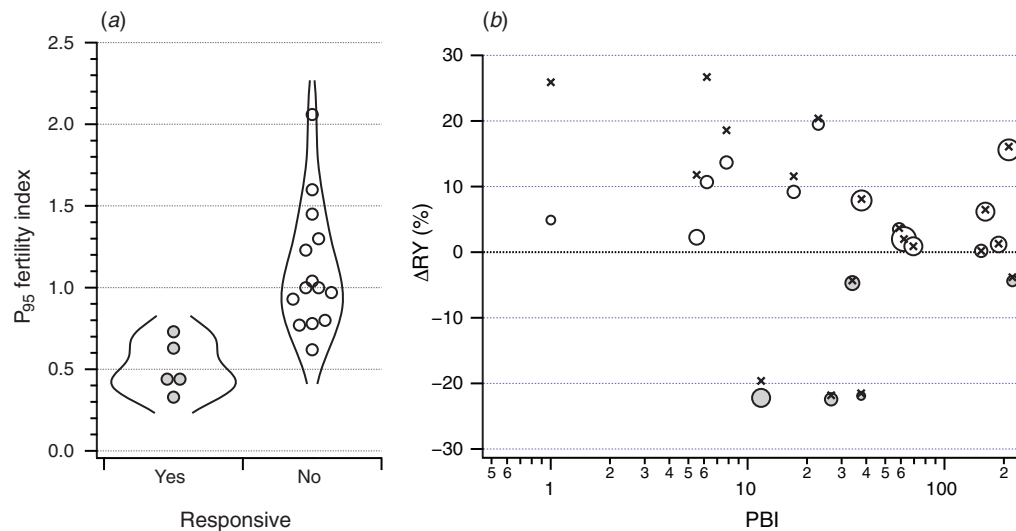


Fig. 3. (a) Violin plots (Hintze and Nelson 1998) of the P_{95} fertility index of the trials classified by whether or not they showed a response to P; and (b) percentage difference in RY (measured – BFDP prediction using Eqn 5 (circles) and Eqn 4 (x)) as a function of soil PBI. Symbol size represents soil pH ranging from 4.1 to 6.2. Actual responsive sites (●); actual non-responsive sites (○).

small dataset of 19 trials to identify such an anomaly. This also highlights the importance of deriving critical soil test values from large datasets such as those in BFDP (Gourley *et al.* 2019), and emphasises that although small datasets such as these 19 trials add value to BFDP, their primary value is to validate.

Differences in RY became systematically wider and more positive when Eqn 4 was used (Fig. 3b), implying that RY would be underestimated by using this equation when PBI is <15. The RY differences assessed by using the equations from Gourley *et al.* (2019) indicate that Eqn 5 better estimates the trial results from this study.

The trial results are consistent with the notion that if soil tests for Colwell P indicate adequate P for pasture production, adding more P does not increase productivity and may increase fertiliser costs and add to the risk of nutrient loss offsite (Weaver and Wong 2011; Gourley *et al.* 2019). Equally, the results indicate that applying P according to soil test should result in the predicted RY. The results also indicate that correcting nutrient deficiencies consistent with the Sprengel–Liebig law of the minimum (van der Ploeg *et al.* 1999) can dramatically increase production (Fig. 2). In the examples of responsive and non-responsive sites shown here, production increases of 137–267% were observed when N, K and S were applied to non-limiting levels. For the responsive site (Fig. 2a), an almost 4-fold increase in dry matter was observed when all nutrient constraints (N, P, K, S) were corrected, again consistent with the Sprengel–Liebig law of the minimum.

The Sprengel–Liebig law of the minimum is clearly shown when the application of basal nutrients led to large increases in dry matter for the same P supply, whether that be from the soil alone (0 kg P/ha), or from the soil and applied P (40 kg P/ha) (Fig. 2). The law of diminishing returns is also demonstrated in Fig. 2a where increases in dry matter decreased with

increasing P application rate, and in Fig. 2b where no further increases in dry matter resulted from P application because there was already sufficient P in the soil. In the case of these trials where P was at or above optimum, large increases in dry matter were related to removal of other nutrient constraints (N, K, S); however, increases or decreases in absolute dry matter yield can also occur through fluctuations in rainfall, highlighting the importance of RY in the determination of soil test critical values. Critical soil test values are unlikely to be related to absolute yield, but are clearly influenced by soil PBI and are related to RY (Gourley *et al.* 2019).

The median P concentration of subterranean clover sampled from the typical responsive trial at 0 kg P/ha with and without basal nutrients was low (0.13 and 0.14%), and would be classified as deficient compared with reported critical values of 0.25–0.5% (Weir and Cresswell 1994), 0.16–0.32% (Sandral *et al.* 2019), and 0.23% (McCaskill *et al.* 2019). For subterranean clover from the same trial when 40 kg P/ha was applied with and without basal nutrients, the P concentration (0.23% and 0.26%) in most cases would be classified as adequate, except when the critical ranges of Reuter and Robinson (1997) are used (0.32–0.70%). The P concentration of ryegrass from the non-responsive trial at 0 kg P/ha with and without basal nutrients (0.26% and 0.31%) was below the critical range described as adequate (0.35–0.70%) by Reuter and Robinson (1997), and would therefore be classified as deficient. Using these same critical ranges, the P concentration of ryegrass from the non-responsive trial for all treatments except when 40 kg P/ha was applied without basal nutrients would be classified as deficient. Compared with the critical range of 0.25–0.55% cited by Weir and Cresswell (1994), the P concentrations of ryegrass in all treatments from the non-responsive trial would be classified as adequate, consistent with the lack of response shown in Fig. 2b. The

critical ranges used to determine deficiency or adequacy in plant tissues require further investigation given that some non-responsive sites showed P concentrations in the plant tissue that could be lower than, or within, critical ranges suggested for adequacy, depending on the range used.

Among these general findings, there were some larger, negative deviations in RY that require further exploration.

Although all of the trials fall within the 95% prediction intervals (Fig. 4), some larger RY deviations of ~20% were found for trials 5, 9 and 11 (Fig. 3*b*). These three sites have soil pH lower than desired ($\text{pH}_{\text{Ca}} 5.5$); however, 16 of the 19 trial sites have low pH. The target value of $\text{pH}_{\text{Ca}} 5.5$ is chosen because the region has almost entirely continuous pasture with minimal soil disturbance when lime is surface-applied (lime

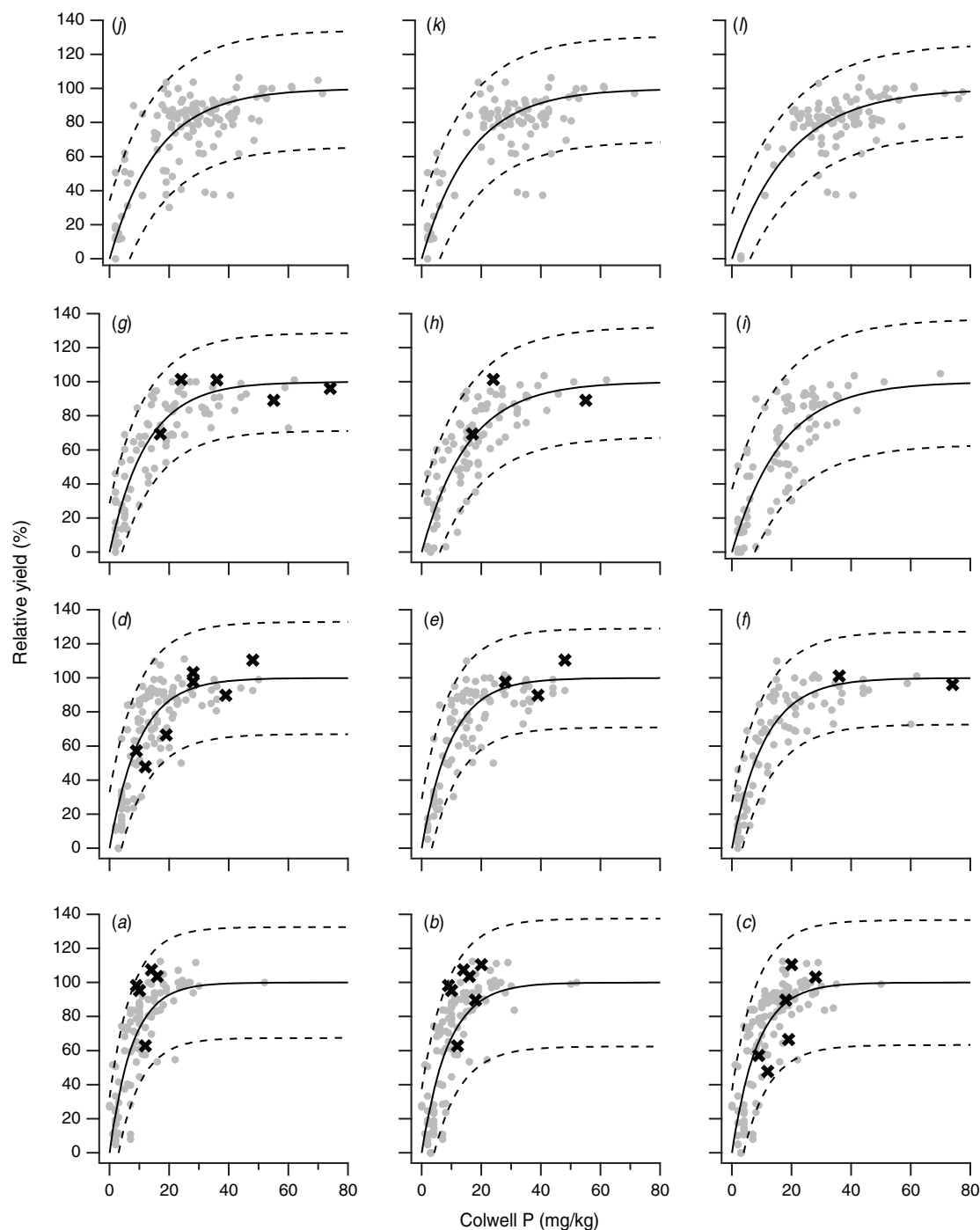


Fig. 4. Response calibrations from BFD for the overlapping PBI ranges specified in Gourley *et al.* (2019). BFD data (●) with fitted response calibration (—) and 95% prediction interval (---), with data from this study overlaid (×). PBI ranges shown (a) –7.2–13.7; (b) 0.9–25.3; (c) 12.6–48.6; (d) 24.2–95.1, (e) 47.4–106.7, (f) 94–164.9, (g) 105.6–223, (h) 163.7–269.5, (i) 221.9–339.3, (j) 268.4–500.6, (k) 338.1–500.6, (l) 499.4–2798.8.

was not applied as part of these trials). Aiming for pH 5.5 in the 0–10 cm layer encourages increases in pH beyond a depth of 10 cm under these conditions, meeting the pH requirements of most pasture species and for nodulation of clover. However, sites 5, 9 and 11 also had low P fertility (Table 3). Observations of the effects of soil pH on the P concentration of subterranean clover provide a possible explanation for the RY deviations (Weaver *et al.* 2020). Those authors observed that the P concentration of clover was significantly depressed under conditions of low P fertility (P_{90} fertility index <0.6) and low pH (pH_{Ca} <5) compared with conditions of similar P fertility but higher pH. Weaver *et al.* (2020) compared critical P ranges for subterranean clover (from Reuter and Robinson 1997) with the P concentrations in subterranean clover sampled from soils with varying P fertility and pH. The authors identified that clover at sites with low P fertility was deficient in P (0.15–0.25% interquartile range) at pH_{Ca} <5 and contained 0.25–0.30% P (interquartile range) at pH_{Ca} >5. Low soil pH is therefore likely to contribute to lower than expected dry matter and RY at some sites. Weaver *et al.* (2020) also observed that the negative effect of soil pH on P concentration in subterranean clover was overcome if the soil had higher P fertility than required as determined by published critical Colwell P values (Gourley *et al.* 2019). Acidic conditions can reduce root growth through aluminium toxicity, limiting the volume of soil from which plants can access P. Increasing soil P fertility could reduce this limitation, because, as indicated by Barrow *et al.* (2020), P availability may not be as limited by low pH as previously thought. Ideally application of lime could ameliorate this issue, reducing downstream consequences of P loss from these paddocks. However, where lime is not applied, Weaver *et al.* (2020) suggested using a P_{90} fertility index of 1.2 for soils with pH_{Ca} <5, above which there was sufficient P in the soil to overcome effects of soil acidity on reduced P uptake. In practice the environmental impact of acidic and low-P sites is limited by the small proportion of such sites; Table 1 indicates that a high proportion of sites in the study area have high P fertility (56% with a P_{95} fertility index >1) and low pH (80% with pH <5.5) and are therefore unlikely to have pasture productivity reduced because of low pH affecting P uptake.

Pasture composition is another factor that may influence RY, pasture responsiveness and the deviations shown in Fig. 3b. Many of the trials conducted here were sown with mixed, contemporary pasture species, and these will not necessarily have the same responsiveness as those used in pasture trials collated by BFDP from 1955 to 2006. For example, it is widely accepted that clover has a higher P requirement than ryegrass (Ozanne *et al.* 1969; Helyar and Anderson 1971; Jackman and Mouat 1972; Barrow 1975; Ozanne *et al.* 1976; Hill *et al.* 2010; Sandral *et al.* 2019). Hence, pasture composition in particular trials and in treatments within trials can also influence responsiveness and, potentially, RY measurements. For example, for the responsive trial (Fig. 2a) where P supply was sufficient (40 kg P/ha) without basal nutrient applied, plots were dominated by 70% clover because the grasses did not have ready access to available N, whereas for the same 40 kg P/ha treatment with basal nutrients applied, plots were dominated

by 90% ryegrass (Goodman and Collison 1981). Notwithstanding these factors, the results here are consistent with the values and variability of data from BFDP.

A simple observation from these trials is the importance of soil testing to providing evidence-based fertiliser decision making. Weaver and Wong (2011), Barrow (2015) and Crawford *et al.* (2020) note the requirement in many situations in Australia for P management to move into a maintenance phase, based around soil testing to guide any capital or maintenance P requirements. This approach needs to replace the traditional approach to P management, which in many cases is to apply P annually irrespective of need, and which has led to legacy P stores that now contribute unnecessarily to offsite water quality problems (Gourley and Weaver 2012). This provides significant opportunity for landholders to redirect current P expenditure to other aspects of the Sprengel–Liebig law of the minimum (van der Ploeg *et al.* 1999), concurrently increasing production and minimising offsite impacts.

Further research is required to establish the economic requirement of targets for RY. Current agronomic guidance is dominated by the dairy industry, which has been at the forefront of soil testing but has a greater yield requirement targeting predominantly 95% of RY for most paddocks. This target is unlikely to be economic or environmentally desirable for the majority of the grazing areas of WA, which are dominated by beef production. This deserves detailed consideration considering the large improvements in yield from non-P inputs shown here.

Conflicts of interest

The authors declare no conflicts of interest.

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