

Phenological Patterns Differ between Exotic and Native Plants: Field Observations from the Sapphire Mountains, Montana

Authors: Durham, Rebecca A., Mummey, Daniel L., Shreading, Lauren, and Ramsey, Philip W.

Source: Natural Areas Journal, 37(3) : 361-381

Published By: Natural Areas Association

URL: <https://doi.org/10.3375/043.037.0310>

The BioOne Digital Library (<https://bioone.org/>) provides worldwide distribution for more than 580 journals and eBooks from BioOne's community of over 150 nonprofit societies, research institutions, and university presses in the biological, ecological, and environmental sciences. The BioOne Digital Library encompasses the flagship aggregation BioOne Complete (<https://bioone.org/subscribe>), the BioOne Complete Archive (<https://bioone.org/archive>), and the BioOne eBooks program offerings ESA eBook Collection (<https://bioone.org/esa-ebooks>) and CSIRO Publishing BioSelect Collection (<https://bioone.org/csiro-ebooks>).

Your use of this PDF, the BioOne Digital Library, and all posted and associated content indicates your acceptance of BioOne's Terms of Use, available at www.bioone.org/terms-of-use.

Usage of BioOne Digital Library content is strictly limited to personal, educational, and non-commercial use. Commercial inquiries or rights and permissions requests should be directed to the individual publisher as copyright holder.

BioOne is an innovative nonprofit that sees sustainable scholarly publishing as an inherently collaborative enterprise connecting authors, nonprofit publishers, academic institutions, research libraries, and research funders in the common goal of maximizing access to critical research.

Phenological Patterns Differ between Exotic and Native Plants: Field Observations from the Sapphire Mountains, Montana

Rebecca A. Durham^{1,2}

¹MPG Ranch
1001 S. Higgins Ave. Suite A3
Missoula, MT 59801

Daniel L. Mummey¹

Lauren Shreading¹

Philip W. Ramsey¹

² Corresponding author:
rdurham@mpgranch.com; 406-360-8155

Natural Areas Journal 37:361–381

ABSTRACT: We collected plant phenology data in the Sapphire Mountains, Montana, USA, by monitoring developmental stages of 101 native species and 21 exotic species during weekly visits to the sites from March to November 2013. We compared the start, end, and length of the emergence, flowering, and seed maturation phases for exotic and native plants. Short-lived forbs, perennial forbs, and perennial grasses were analyzed separately. Exotic plants emerged earlier, began and ended flowering later, and had later ends to emergence and dispersal phases across all functional group comparisons. The emergence phase for exotic perennial forbs averaged 13.8 weeks longer than for native perennial forbs, the flowering phase was 3.4 weeks longer, and the seed dispersal phase was 8.3 weeks longer. The window for emergence and flowering for forbs, shrubs, and grasses (March to November) did not differ between natives and exotics. The results generally support the conclusion that the exotics have an advantage over the natives in priority of growth and wider niche breadth, rather than occupying vacant niches. Seed set time varied in duration from 1 to 23 weeks from April to November. Our results provide insights into invasion mechanisms and selection of native plant materials to compete with invasive species. We discuss implications for seed collection and herbicide timing.

Index terms: herbicide timing, microsite variation, native seed collection, phenology, plants

INTRODUCTION

Plant phenology, the seasonal timing of plant developmental stages, influences plant community dynamics and determines resource availability to herbivores and pollinators. Increased attention has been directed to differences in native and exotic plant phenology and how those differences affect invader abundance (Godoy et al. 2009b; Wolkovich and Cleland 2010; Wainwright et al. 2012). Specific exotic plants have been shown to have earlier or later flowering periods (Godoy et al. 2009a, 2009b; Pearson et al. 2012) and earlier germination than native species (Wainwright et al. 2012; Wainwright and Cleland 2013), but the generality of these observations has not been investigated.

The geographic origin of plant species influences phenology, as plants have evolved to the climatic conditions of their native range (Rathke and Lacey 1985; Godoy et al. 2009a). Though plants can change through selective pressure, they often retain the phenological patterns of their original range when they occupy a new range (Franks et al. 2007; Maron et al. 2007; Godoy et al. 2009a; Lesica and Kittelson 2010). Life history strategy also creates disparate phenological patterns (Jia et al. 2011). For example, annual grasses will allocate more resources to reproduction than perennial grasses that must invest in maintenance for longevity.

While most studies focus on one particular temporal phenological stage, usually the

flowering stage, none have focused on how lengths of all stages vary through a growing season. Timing and duration trends of all plant phenological stages may inform us how exotic invaders compete with and displace natives. For example, a longer period of flowering and seed dispersal aids genetic mixing in invasive populations and allows the invader to take advantage of a wider array of favorable conditions across a growing season (Cadotte et al. 2006). A longer length of emergence period could give greater access to limited resources and, thus, increase fecundity (Cadotte and Lovett-Doust 2002).

We measured phenological traits for an entire growing season to test whether exotic species differ in timing and phase length compared to native species and whether these differences occur at the species functional group level. We considered exotic species in this study as any alien-origin plant species, whether it was invasive, noxious, or benign. Along with collecting species-level phenology in communities where nonnative species threaten the integrity of ecosystem function, an aim of this study was to understand plant phenology in regards to restoration of native plant communities in degraded systems. Though research has highlighted the importance of plant phenology for management applications (Ghersa and Holt 1995; Ansquer et al. 2009; Pearson et al. 2012), a gap remains in the knowledge of phenological patterns of native and exotic plants in the northern Rocky Mountains. For example, restoration aimed at filling

niche space to best use available resources must account for both spatial and temporal plant relationships. Restoration maximizing niche space requires knowledge of phenological differences. Also, there is a notion among managers that windows exist for herbicide applications when damage to native plant populations can be minimized because problematic exotic plants are active aboveground while most native plants are dormant (Rice et al. 1997). These opportunities are referred to as “management windows” and would occur in cases where exotics occupied a vacant temporal niche in the community (Wolkovich and Cleland 2010). We found no comprehensive phenology data to support the existence of management windows, but we evaluate this concept.

METHODS

Study Site

This study was conducted at MPG Ranch in the Northern Sapphire Mountains of western Montana (Table 1, Figure 1). The ranch is a 3800-ha conservation property with topography that varies from flat bottomland to gentle foothills and forested mountain slopes. A combination of agricultural land, sagebrush steppe, riparian forests, dry open forests, and moist mixed coniferous forests dominate the landscape. Mean annual precipitation ranges from 300 mm on the valley floor to 350 mm on mountain summits, and mean temperatures in nearby Missoula for July and January are 19.4 °C and -4.7 °C, respectively. Cold moist winters, when most precipitation occurs, contrast with hot, dry summers.

Phenology study areas centered on 15 sites representative of habitats in the lower to upper montane rangeland (Table 1). Each area encompassed approximately 730 m². Native species richness ranged from 7 to 40 species across the sites. Exotic species, mainly *Centaurea stoebe* L. (spotted knapweed), *Euphorbia esula* L. (leafy spurge), *Bromus tectorum* L. (cheatgrass), *Sisymbrium altissimum* L. (tumble mustard), and *Potentilla recta* L. (sulfur cinquefoil) comprised 1%–84% of the plant cover (% cover from point-in-

Table 1. Survey site characteristics.

Site	Coordinates		Elevation (m)	Ecological system	Number of species per plot		
	Latitude	Longitude			Native	Exotic	Total
1	46.68605°	-113.98506°	1308	Rocky Mountain Montane-Foothill Deciduous Shrubland	36	14	50
2	46.70457°	-114.00040°	1480	Rocky Mountain Lower Montane, Foothill, Valley Grassland	29	6	35
3	46.70369°	-114.00676°	1325	Montane Sagebrush Steppe	26	3	29
4	46.70436°	-114.00352°	1386	Montane Sagebrush Steppe	41	8	49
5	46.69840°	-114.01187°	1233	Montane Sagebrush Steppe	34	9	43
6	46.69531°	-114.01762°	1172	Montane Sagebrush Steppe	4	7	11
7	46.68076°	-113.99016°	1267	Montane Sagebrush Steppe	10	8	18
8	46.67767°	-113.99591°	1189	Montane Sagebrush Steppe	14	11	25
9	46.67811°	-113.99272°	1223	Montane Sagebrush Steppe	11	8	19
10	46.71383°	-113.99976°	1491	Montane Sagebrush Steppe	18	9	27
11	46.71152°	-113.99572°	1468	Rocky Mountain Dry-Mesic Montane Mixed Conifer Forest	40	8	48
12	46.70501°	-113.98061°	1748	Rocky Mountain Subalpine-Upper Montane Grassland	28	4	32
13	46.70722°	-113.98124°	1817	Rocky Mountain Subalpine-Upper Montane Grassland	32	3	35
14	46.71428°	-113.99657°	1539	Rocky Mountain Lower Montane, Foothill, Valley Grassland	18	6	24
15	46.71472°	-113.99338°	1544	Rocky Mountain Lower Montane, Foothill, Valley Grassland	21	6	27

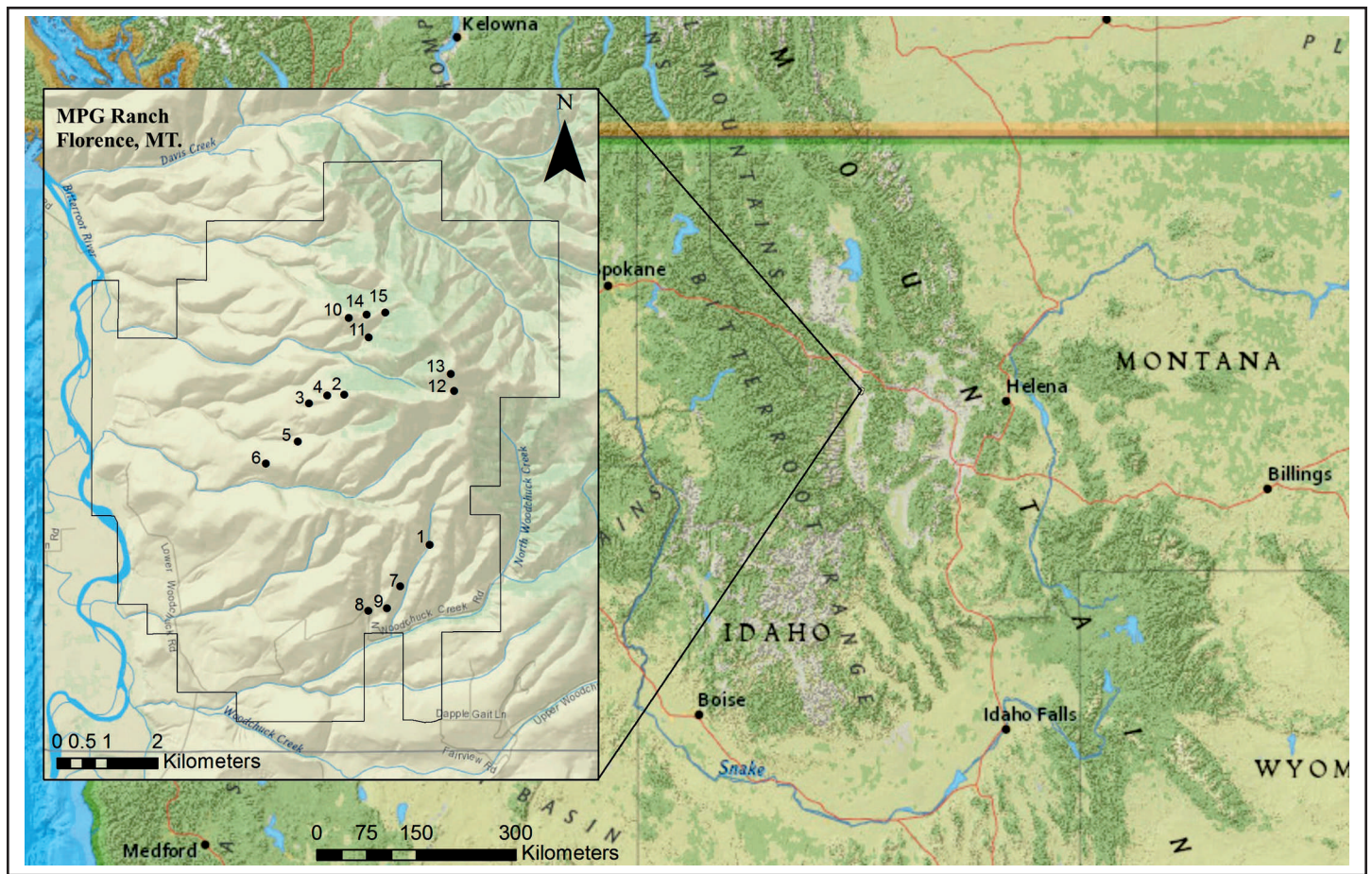


Figure 1. Study site locations within the boundary of MPG Ranch, Montana, USA.

tercept transect data, unpublished). Site elevation ranged from 1172 m to 1817 m (Table 1). Soil temperature and weather data can be accessed from <http://livecams.mpgranch.com>.

Field Sampling

We assigned each species encountered at each site to the category (or categories) emergent, budding, flowering, fruiting, mature seed, senesced, and fall growth/fall germination, weekly from 10 March 2013 to 23 November 2013. We visited sites the same day each week (± 2 days). Phases were not mutually exclusive; we recorded each phase presented per species by site. Phenological stage was considered present if at least 80% of the individuals at the site were in the stage, to minimize noise from intraspecific variation.

Plants with green tissue were recorded as

emergent. Budding plants had immature reproductive structures and closed sepals. Flowering plants were considered those with mature bloom and developed floral structures (i.e., anthers, stigma, petals). The fruiting period was recorded as the period with fruit after the flower withered, regardless of fruit maturity. Mature seed was recorded for fruiting plants with mature, collectable seed. We recorded plants with withered non-green vegetation as senesced. Fall regrowth was assigned to those plants that showed new vegetative growth in the fall. We recorded fall germination for plants that sprouted new seedlings in the fall.

Data Analysis

Homogeneity of variance within functional groups and differences in the number of native and exotic species in functional groups presented a challenge to analysis. For example, there were 60 native peren-

nial forbs represented at three or more study sites. Only four exotic perennial forbs were detected (*C. stoebe*, *E. esula*, *Silene latifolia* Poir. [bladder campion], and *P. recta*). *Centaurea stoebe*, *E. esula*, and *P. recta* were well represented, but *S. latifolia* was only present at one site and was not observed in all stages. Differences in sample numbers between plant functional groups caused the data to violate the ANOVA assumptions of normality and homogeneity of variance. This problem could not be corrected by data transformations. For these reasons, we analyzed the data using nonparametric methods. We applied a Kruskal–Wallis *H* test to determine differences in duration and timing of phenological phases between functional groups (short-lived or perennial; forb or graminoid) and species origin (native or exotic). When species occurred at more than one site, average values were used to minimize noise across sites. A Mann–Whitney *U* test was used to test

differences between subgroups. We discuss the phenology of plants within functional groups in relation to the distribution of the native plants within the same functional group. These analyses were performed with SPSS version 20.

RESULTS

Flowering Phenology

Flower duration, the number of weeks a species was in bloom, was charted for each study species by life form (Appendix). Flower duration for each species was the first week of occurrence to the last week of occurrence across all sites. Short-lived forbs included annuals and biennials. Perennial native forbs and native short-lived forbs flowered from March to September (Appendix). Exotic perennial forbs flowered from May to November, and exotic short-lived forbs flowered from March to November. Duration of flowering for native forbs was 2–18 weeks depending on the species, and flowering duration for exotic forbs was 2–31 weeks depending on the species (Appendix). Native perennial graminoids flowered from the end of April to the beginning of August. Exotic perennial graminoids flowered from the end of April to the beginning of July (Appendix). Duration of flowering for native perennial graminoids was 1–10 weeks and flowering duration of exotic perennial graminoids was 1–9 weeks (Appendix). Native shrubs flowered from the end of April to the beginning of November. Flower duration for shrubs was 3–13 weeks. Exotic annual grasses flowered from May to June, flower duration was 1–7 weeks (Appendix).

Phenology of Native vs. Exotic Functional Groups

We compared growth phases of the three functional groups with native and exotic representatives: short-lived forbs, perennial forbs, and perennial graminoids. Exotic and native species exhibited disparate phenology for the periods of emergence, flowering, and mature seed (Figures 2 and 3), while budding, fruiting, senescence, fall growth, and fall emergence did not differ among groups and are not reported. The

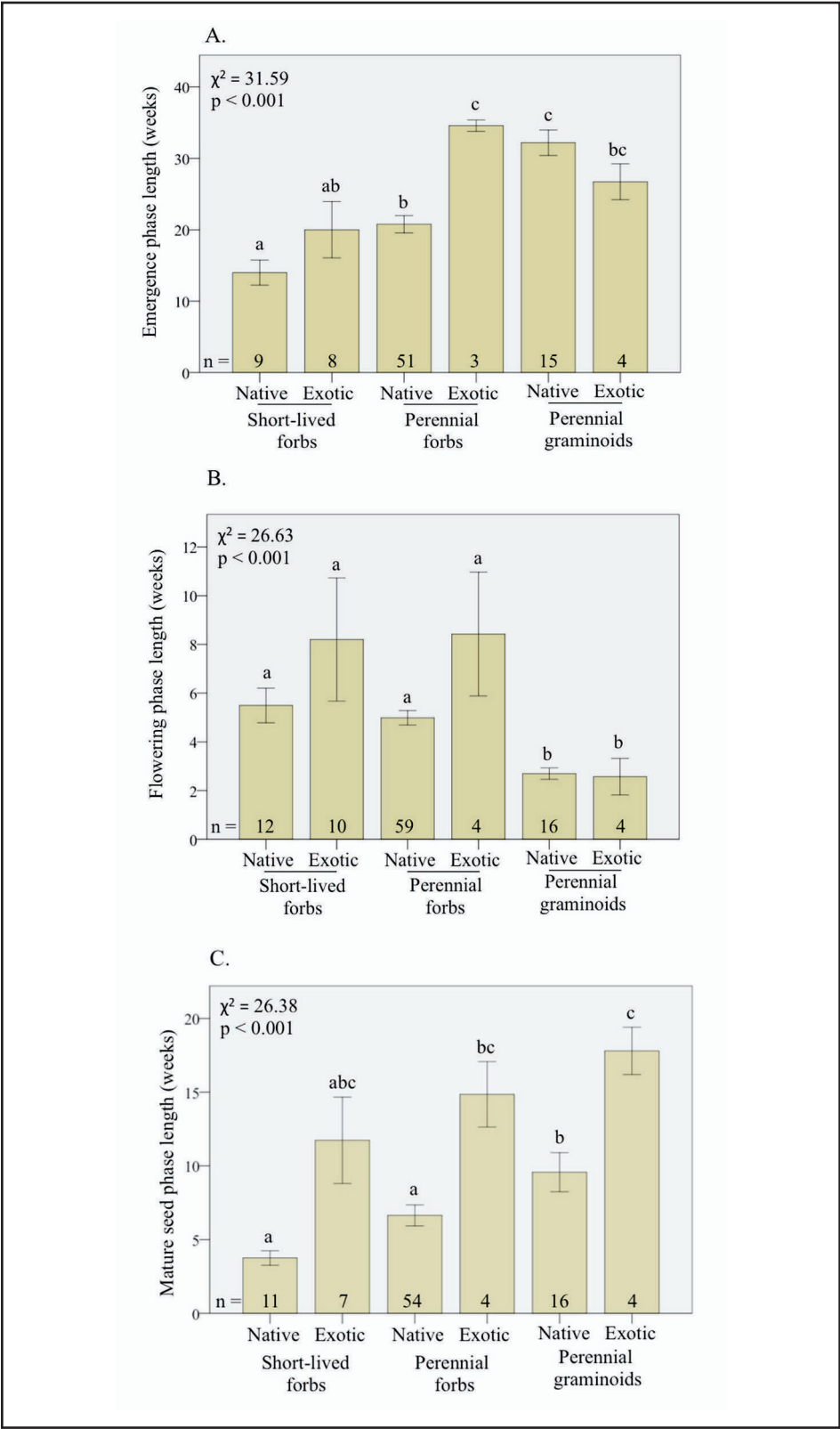


Figure 2. Mean duration of phenology phases of native and exotic functional groups: short-lived forbs, perennial forbs, and perennial graminoids (+/- 1 SE). Different letters indicate significant differences ($P < 0.05$). Phenology phases represented are (A) emergent phase length, the period of time when plant species have green plant tissue; (B) flowering phase length, the period of time when plant species have developed floral structures; and (C) mature seed phase length, the period of time when fruiting plant species have mature, collectible seeds.

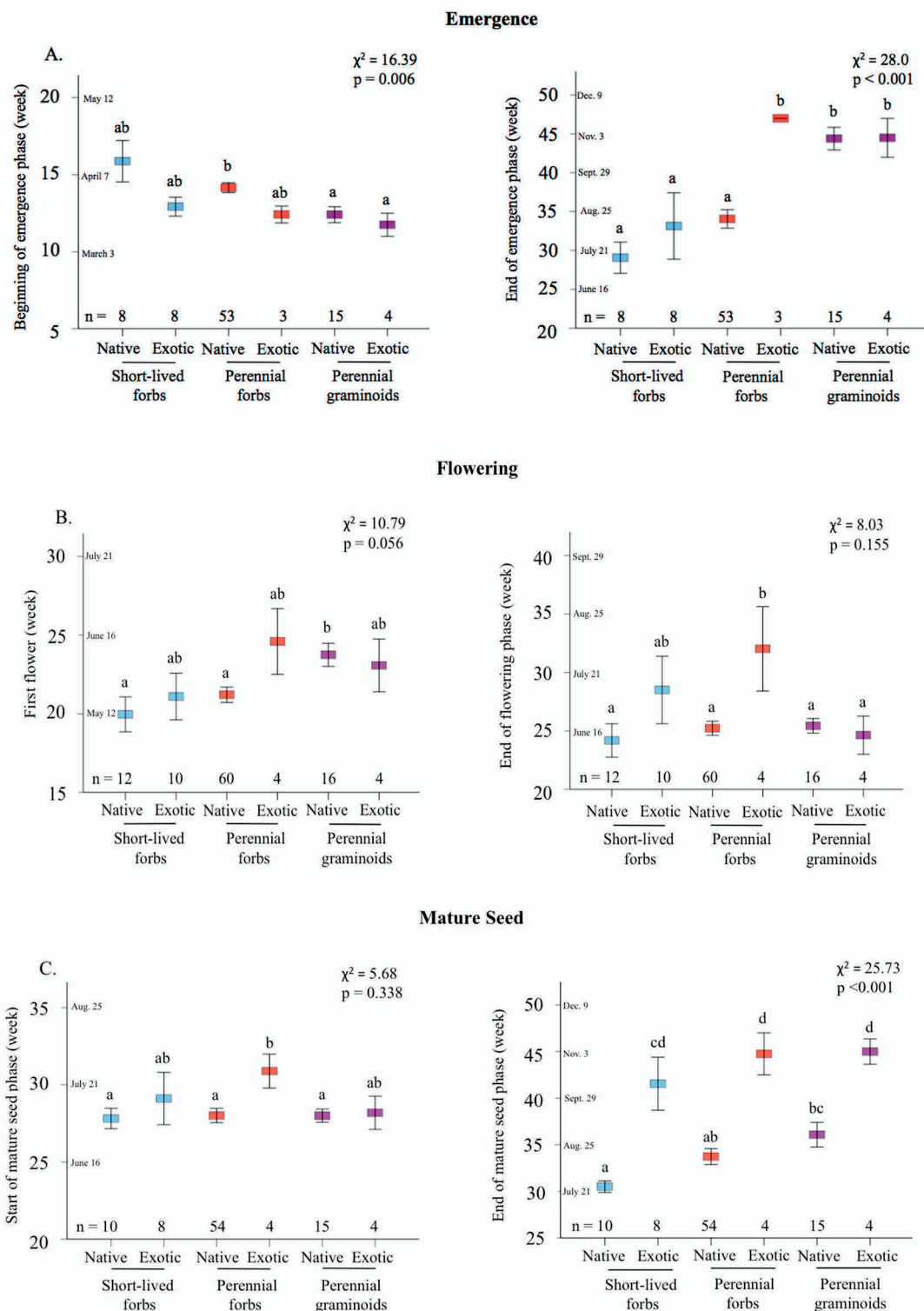


Figure 3. Mean timing (phase start and phase end) of phenology phases of native and exotic functional groups: short-lived forbs, perennial forbs, and perennial graminoids (± 1 SE). Different letters indicate significant differences ($P < 0.05$). Phenology phases represented are (A) the start and end of the emergence phase, when plant species have green plant tissue (start) and when green tissue is gone (end); (B) the start and end of the flowering phase, when plant species develop the first floral structures (start) and when the last floral structures are gone (end); and (C) the start and end of the mature seed phase, when plant species have the first mature, collectible seeds (start) and when the last mature seeds drop (end).

rest of this section will discuss differences between exotic and native short-lived forbs, perennial forbs, and perennial graminoids.

Short-Lived Forbs

The emergence phase length and beginning and end of the emergence phase were sim-

ilar for native and exotic short-lived forbs (Figures 2 and 3). Flowering periods and the beginning and end of the flowering periods were similar for native and exotic short-lived forbs (Figures 2 and 3). The biggest differences between native and exotic short-lived forbs were noted for the mature seed phase. The exotic short-lived forbs trended toward holding mature

seed longer than natives ($Z = -1.87$, $P = 0.069$; Table 2, Figure 2). The mature seed phase ended significantly later for exotic short-lived forbs than natives ($Z = -2.38$, $P = 0.016$; Table 2, Figure 3). Most of the exotic short-lived forbs had mature seed for most of the growing season, 11.7 weeks ($SD \pm M7.77$). The native short-lived forbs held seed for a window

Table 2. Phase lengths, beginning and end week for the phenological phases of short-lived forbs, perennial forbs, and perennial graminoids.

Phenological phase	<i>n</i>		Mean (standard deviation)		Mann-Whitney <i>U</i>	
	native	exotic	native	exotic	<i>Z</i>	<i>p</i>
Short-lived forbs						
Emergence						
Length (weeks)	9	8	14.0 (5.24)	20.0 (11.2)	-1.3	0.2
Start (week)	8	8	15.9 (3.80)	12.9 (1.74)	-1.49	0.161
End (week)	8	8	29.1 (5.69)	33.1 (12.1)	-0.158	0.878
Flowering						
Length (weeks)	12	10	5.49 (2.45)	8.20 (8.00)	-0.33	0.771
Start (week)	12	10	20.0 (3.85)	21.1 (4.67)	-0.4	0.72
End (week)	12	10	24.2 (4.97)	28.5 (9.11)	-1.45	0.159
Mature seed						
Length (weeks)	11	7	3.75 (1.63)	11.7 (7.77)	-1.87	0.069
Start (week)	10	9	27.8 (2.08)	29.1 (5.09)	-0.9	0.4
End (week)	10	8	30.5 (1.97)	41.6 (8.05)	-2.38	0.016
Perennial forbs						
Emergence						
Length (weeks)	51	3	20.8 (8.64)	34.6 (1.38)	-2.25	0.019
Start (week)	53	3	14.2 (2.32)	12.4 (0.946)	-1.54	0.133
End (week)	53	3	34.1 (8.62)	47.0 (0.000)	-2.31	0.015
Flowering						
Length (weeks)	59	4	4.99 (2.26)	8.42 (5.08)	-1.45	0.146
Start (week)	60	4	21.2 (3.77)	24.6 (4.17)	-1.55	0.127
End (week)	60	4	25.2 (4.69)	32.0 (7.22)	-2.4	0.012
Mature seed						
Length (weeks)	54	4	6.64 (5.27)	14.9 (4.44)	-2.54	0.007
Start (week)	54	4	28.0 (3.43)	30.9 (2.21)	-2.12	0.032
End (week)	54	4	33.7 (6.26)	44.8 (4.50)	-2.68	0.004
Perennial graminoids						
Emergence						
Length (weeks)	15	4	32.2 (6.87)	26.7 (5.00)	-1.6	0.124
Start (week)	15	4	12.4 (2.01)	11.8 (1.50)	-1.08	0.307
End (week)	15	4	44.4 (5.63)	44.5 (5.00)	-0.07	0.961
Flowering						
Length (weeks)	16	4	2.69 (0.938)	2.57 (1.50)	-0.52	0.617
Start (week)	16	4	23.8 (2.94)	23.1 (3.35)	-0.28	0.82
End (week)	16	4	25.4 (2.50)	24.6 (3.27)	-0.57	0.617
Mature seed						
Length (weeks)	16	4	9.57 (5.31)	17.8 (3.21)	-2.22	0.022
Start (week)	15	4	28.0 (1.65)	28.2 (2.15)	-0.65	0.53
End (week)	15	4	35.1 (5.10)	45.0 (2.72)	-2.41	0.014

of less than a month at 3.75 weeks (SD \pm 1.63). *Erodium cicutarium* (L.) L'Her. ex Aiton (storksbill) was one of the plants contributing to this statistic, though other short-lived exotics mirrored this pattern. *Erodium cicutarium* did not hold seed from single flowers through the growing season; instead the plants had cycles of flowering and fruiting that continuously kept mature seed present for 11 weeks.

Perennial Forbs

Three exotic perennial forbs dominate many plant communities across the study area and our region. These are *C. stoebe*, *E. esula*, and *P. recta*. Of these, *C. stoebe* and *E. esula* are the most abundant and

problematic. One other exotic perennial forb, *S. latifolia*, was detected at a study site. Unlike the other three forbs, *S. latifolia* poses little threat to invasion in our rangelands (pers. obs.). Due to the uneven numbers of species in the native versus the exotic perennial forb groups (Table 2), we describe the differences between the exotic and native perennial forbs as a group (Table 2) and individually (Table 3). Overall, the exotic forbs remained active longer (emergence phase $Z = -2.25$, $P = 0.019$; Figure 2), flowered later ($Z = -2.4$, $P = 0.012$; Figure 3), and retained mature seed more than twice as long as the native perennial forbs ($Z = -2.54$, $P = 0.007$; Figures 2 and 3).

Centaurea stoebe's emergent growth phase

lasted more than 14 weeks longer than the mean of the native perennial forbs (Table 3). *Centaurea stoebe*'s flowering phase was 9 weeks longer than the mean of the natives and it carried mature seed for twice as long as the native mean (13.9 weeks for *C. stoebe*, compared to 6.64 weeks for the native perennial forbs; Table 3).

Euphorbia esula's emergent growth phase was also longer than the mean of the native perennial forbs (34.5 weeks; Table 3). The mean of the *E. esula* flowering period was twice as long as the mean of the native perennial forb flowering period (Table 3). *Euphorbia esula* carried mature seed for longer and later than the mean of the natives, but only by about 2.5 weeks (Table 3).

Table 3. Phase lengths, start week, and end week for native perennial forbs and exotic perennial forbs.

	Emergence			Flowering			Mature seed		
	Length (weeks)	Start (week)	End (week)	Length (weeks)	Start (week)	End (week)	Length (weeks)	Start (week)	End (week)
Native perennial forbs									
<i>n</i>	51	53	53	59	60	60	54	54	54
mean	20.80	14.20	34.10	4.99	21.20	25.20	6.64	28.00	33.70
s.d.	8.60	2.30	8.60	2.30	3.80	4.70	5.30	3.40	6.30
Spotted knapweed (<i>Centaurea stoebe</i>)									
<i>n</i>	13	13	13	13	13	13	13	13	13
mean	36.00	12.00	47.00	14.70	31.50	39.50	13.90	34.10	47.00
s.d.	2.00	2.00	0.00	2.20	3.50	3.70	0.28	0.28	0.00
Leafy spurge (<i>Euphorbia esula</i>)									
<i>n</i>	2	2	2	4	4	4	4	4	4
mean	34.50	13.50	47.00	9.50	20.30	28.80	9.00	30.00	38.00
s.d.	0.71	0.71	0.00	0.58	0.96	0.50	1.40	0.82	2.00
Bladder campion (<i>Silene latifolia</i>)									
<i>n</i>	0	0	0	1	1	1	1	1	1
mean	n.d.	n.d.	n.d.	7.00	22.00	28.00	29.00	19.00	47.00
s.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.
Sulfur cinquefoil (<i>Potentilla recta</i>)									
<i>n</i>	4	4	4	4	4	4	4	4	4
mean	33.30	11.80	38.80	2.50	27.00	28.50	17.50	30.50	47.00
s.d.	0.50	1.50	5.60	1.30	1.20	0.58	0.58	0.58	0.00
Sticky cinquefoil (<i>Potentilla glandulosa</i>)									
<i>n</i>	1	1	1	1	1	1	1	1	1
mean	21.00	16.00	36.00	4.00	24.00	27.00	19.00	29.00	47.00
s.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.
Slender cinquefoil (<i>Potentilla gracilis</i>)									
<i>n</i>	n.d.	n.d.	n.d.	1	1	1	1	1	1
mean	n.d.	n.d.	n.d.	3.00	26.00	28.00	9.00	33.00	41.00
s.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.

We monitored only one population of *S. latifolia* in this study and it was not detected until the budding stage, so emergence data is not presented. Anecdotally, *S. latifolia* flowered for about as long as the natives, but held mature seed from the end of the flowering period until the end of the growing season (29 weeks; Table 3).

Differences in phenology between *P. recta* and the native perennial forbs were not as large as between *C. stoebe*, *E. esula*, and the native perennial forbs. *Potentilla recta* was also the only perennial exotic forb with congeneric species in the study: *Potentilla glandulosa* Lindl. (sticky cinquefoil) and *Potentilla gracilis* Douglas ex Hook. (slender cinquefoil). *Potentilla recta*'s period of emergence was longer than the mean of the natives by 12 weeks and longer than the congeneric *P. glandulosa*, which was emerged for a period close to the mean of the native perennial forbs (21 weeks; Table 3). *Potentilla recta* and its congeners all held mature seed for longer than the mean of the natives, with *P. recta* and *P. glandulosa* averaging 17.5 and 19 weeks, respectively (Table 3).

Perennial Graminoids

The exotic perennial grasses in the study area were *Poa bulbosa* L. (bulbous bluegrass), *Thinopyrum intermedium* (Host.) Barkworth & Dewey (intermediate wheatgrass), *Poa pratensis* L. (Kentucky bluegrass), and *Agropyron cristatum* (L.) Gaertn. (crested wheatgrass). Most grasses, and all the exotics, were emerged through the growing season, flowered approximately the same time, and displayed similar phenology throughout the growing season with the only exception being with the time that exotics held mature seed. Exotic graminoids held seed more than 10 weeks longer than natives ($Z = -2.41$, $P = 0.014$; Figure 3).

Mature Seed Collection

Mature seed data for native species are presented in Table 4. When the species occurred at more than one site, we combined mature seed data to show the interval of when mature seed was available for

collection. Mature seed period varied in duration (1–23 weeks) and time (April–November). When species occurred at multiple sites with similar habitats, we charted whether we observed variability in phenological stages between sites. We defined this variability as greater than or equal to two weeks' difference between the start of phenological stage. If we observed phenological variability across sites, the species was suspected to show microsite sensitivity. Microsite sensitivity occurs when species respond to specific environmental cues at a microsite level, such as aspect or shade (Titus and Tsuyuzaki 2003) or when competition between species varies on small scales (Larigauderie and Richards 1994). Some species showed these differences in phenological patterns at every stage across microsites, and others only for certain stages (Table 4).

Peak Emergence and Management Windows

We also examined peak emergence periods for native and exotic grasses and forbs to gauge differences by functional group. We considered peak emergence to be when most species were emergent with any green tissue across sites. The peak emergent period for native forbs and shrubs was from late May to early June, and the period of least activity was at the beginning and end of the growing season, early March and mid-October to November (Figure 4). The peak emergent period of exotic forbs was April to early May and the period of least activity was March and late October through November. The period when exotic forb emergence was highest and native forb emergence was lowest was late March and mid-October to November. The peak emergent period of native grasses was early May to late June, but remained high throughout the growing season (Figure 4). The peak emergent period of exotic grasses was mid-April to early June, and period of least activity was early to mid-August to mid-September (Figure 4). The period when exotic grass presence was highest and native forb presence was lowest was in early March (Figure 4).

DISCUSSION

Phenology of Native and Exotic Plants

Our study presents an extensive account of phenological stages for 101 native species and 21 exotic species through one growing season at a western Montana rangeland. In the plant communities studied here, exotic perennial forbs have longer periods of emergence than native perennial forbs. We found that in each functional group exotics have longer periods of mature seed and a later end of dispersal period than natives. Exotic forbs tend to emerge earlier, begin and end flowering later, and their dispersal phase tended to start later.

We divided plants into functional groups for comparisons. Though the number of exotic perennial forbs in the sample was only four, these plants are abundant across the study sites. A longer window of emergence, flowering, and seed dispersal gives the exotic perennial forbs greater access to resources, pollinators, and a longer window for seed to move to new locations. For example, some of the trends of the perennial forb functional group were driven by *C. stoebe*, whose phenology was very different from most native perennial plants. For *C. stoebe*, the flowering and mature seed phases happened concurrently, and were longer and later than for native perennial forbs. This extended cycle of flowering, fruiting, and seeding could allow for more genetic mixing between and within populations as different pollinators and dispersers move through the system. At the beginning and end of the flowering period for *C. stoebe*, only a few flowers were found in the populations at the study sites. This forces pollinators to move long distances between flowers, facilitating long-distance pollen dispersal. During the height of the flowering period, flowers are abundant, ensuring that pollinators do not need to travel far and the plants within a local population will be genetically well mixed (Godoy et al. 2009a). Flowering when most natives are fruiting reduces pollinator competition and allows for more pollination success for exotics (Godoy et al. 2009b; Pearson et al. 2012).

Table 4. The window of time when mature seed was collectible from native study species and the variability of all phenology phases across sites. The microsite sensitivity, when phenology phase start differed across sites by more than two weeks (yes or no). If microsite sensitivity was observed for a particular phase, the phase is noted: EM: emergent phase, BU: budding phase, FL: flowering phase, FR: fruiting phase, MS: mature seed phase, SE: senescent phase.

Scientific name	Common name	Life form	Date of first mature seed	Duration of mature seed (weeks)	Observed microsite sensitivity (≥2 weeks)	Observed phase variability
<i>Acer glabrum</i>	Rocky Mountain maple	shrub	18-Aug	14	N/A	
<i>Achillea millefolium</i>	western yarrow	forb	11-Aug	15	Yes	EM, BU, FR, MS SE
<i>Agastache urticifolia</i>	giant horsemint	forb	21-Jul	10	Yes	EM, BU, FL
<i>Agoseris glauca</i>	false-dandelion	forb	9-Jun	2	N/A	
<i>Allium cernuum</i>	nodding onion	forb	21-Jul	18	Yes	BU, FL
<i>Amelanchier alnifolia</i>	saskatoon	shrub	7-Jul	11	N/A	
<i>Antennaria dimorpha</i>	low pussytoes	forb	28-Apr	6	N/A	
<i>Antennaria rosea</i>	rosy pussytoes	forb	9-Jun	9	Yes	FR, MS
<i>Antennaria umbrinella</i>	umber pussytoes	forb	9-Jun	9	Yes	FL, FR, MS
<i>Arenaria congesta</i>	ballhead sandwort	forb	30-Jun	8	Yes	EM, BU, FL, FR, MS, SE
<i>Aristida purpurea</i>	three-awn	graminoid	14-Jul	19	No	
<i>Arnica sororia</i>	twin arnica	forb	16-Jun	9	Yes	FL, BU, MS, SE
<i>Artemisia dracunculus</i>	wild tarragon	shrub	10-Nov	2	No	
<i>Artemisia frigida</i>	fringed sagebrush	shrub	3-Nov	3	No	
<i>Artemisia tridentata</i>	big sagebrush	shrub	3-Nov	3	No	
<i>Astragalus inflexus</i>	hairy milkvetch	forb	21-Jul	2	No	
<i>Astragalus miser</i>	timber milkvetch	forb	30-Jun	3	Yes	FR, SE
<i>Balsamorhiza sagittata</i>	balsamroot	forb	16-Jun	9	Yes	EM, BU, FL, FR, MS, SE
<i>Bouteloua gracilis</i>	blue grama grass	graminoid	21-Jul	6	Yes	MS
<i>Bromus carinatus</i>	mountain brome	graminoid	7-Jul	16	Yes	EM
<i>Carex filifolia</i>	threadleaf sedge	graminoid	7-Jul	6	N/A	
<i>Carex geyeri</i>	elk sedge	graminoid	16-Jun	5	N/A	
<i>Carex petasata</i>	Liddon's sedge	graminoid	7-Jul	7	No	
<i>Castilleja hispida</i>	harsh paintbrush	forb	14-Jul	10	N/A	
<i>Chrysothamnus viscidiflorus</i>	green rabbitbrush	shrub	15-Sep	10	No	
<i>Clarkia pulchella</i>	pink fairies	forb	28-Jul	3	N/A	
<i>Collinsia parviflora</i>	blue-eyed Mary	forb	16-Jun	8	Yes	EM, FL, FR, MS, SE
<i>Collomia linearis</i>	narrow-leaved collomia	forb	7-Jul	6	Yes	FR
<i>Crataegus douglasii</i>	black hawthorn	shrub	21-Jul	4	N/A	

Continued

Table 4. (Cont'd)

Scientific name	Common name	Life form	Date of first mature seed	Duration of mature seed (weeks)	Observed microsite sensitivity (≥2 weeks)	Observed phase variability
<i>Crataegus douglasii</i>	black hawthorn	shrub	21-Jul	4	N/A	
<i>Crepis atriobarba</i>	slender hawksbeard	forb	23-Jun	2	N/A	
<i>Crepis intermedia</i>	gray hawksbeard	forb	30-Jun	3	No	
<i>Danthonia unispicata</i>	onespike oatgrass	graminoid	21-Jul	6	N/A	
<i>Delphinium bicolor</i>	larkspur	forb	23-Jun	3	Yes	EM, BU, FL, FR
<i>Dodecatheon conjugens</i>	Bonneville shootingstar	forb	16-Jun	4	Yes	EM, BU, FL, FR
<i>Dodecatheon pulchellum</i>	few flowered shooting star	forb	16-Jun	7	Yes	MS, SE
<i>Draba nemorosa</i>	woods draba	forb	21-Jul	1	N/A	
<i>Elymus glaucus</i>	blue wild rye	graminoid	28-Jul	17	N/A	
<i>Ericameria nauseosa</i>	gray rabbitbrush	shrub	15-Sep	10	Yes	BU, FL
<i>Erigeron compositus</i>	cutleaf daisy	forb	19-May	4	Yes	FL, FR, MS
<i>Erigeron divergens</i>	spreading fleabane	forb	23-Jun	12	Yes	BU, FL, FR, MS, SE
<i>Erigeron pumilus</i>	shaggy fleabane	forb	30-Jun	4	Yes	MS
<i>Erigeron speciosus</i>	aspen fleabane	forb	21-Jul	4	Yes	FR
<i>Eriogonum ovalifolium</i>	cushion buckwheat	forb	14-Jul	5	N/A	
<i>Eriogonum umbellatum</i>	sulphur buckwheat	forb	14-Jul	5	Yes	BU, FL
<i>Festuca campestris</i>	rough fescue	graminoid	30-Jun	7	Yes	EM
<i>Festuca idahoensis</i>	Idaho fescue	graminoid	23-Jun	8	Yes	EM, FL, FR, MS, SE
<i>Fritillaria pudica</i>	yellow bell	forb	30-Jun	5	Yes	EM, BU, FL, FR
<i>Gaillardia aristata</i>	blanket flower	forb	7-Jul	20	Yes	BU, FR, MS, SE
<i>Geranium viscosissimum</i>	sticky geranium	forb	14-Jul	4	Yes	SE
<i>Geum triflorum</i>	prairie smoke	forb	16-Jun	8	Yes	EM, FR, MS, SE
<i>Heterotheca villosa</i>	hairy golden aster	forb	14-Jul	12	Yes	EM, FR, SE
<i>Hieracium scouleri</i>	Scouler's woollyweed	forb	21-Jul	2	Yes	EM, BU, SE
<i>Hydrophyllum capitatum</i>	ball head waterleaf	forb	16-Jun	6	Yes	FL
<i>Koeleria macrantha</i>	June grass	graminoid	7-Jul	9	Yes	EM, BU, FL, FR, MS
<i>Lewisia rediviva</i>	bitterroot	forb	30-Jun	3	Yes	EM, FL, MS
<i>Lithophragma glabrum</i>	smooth fringe cup	forb	23-Jun	10	N/A	
<i>Lithophragma parviflorum</i>	smallflower woodland star	forb	16-Jun	11	Yes	EM, BU, FL, FR, SE
<i>Lithospermum ruderale</i>	field gromwell	forb	30-Jun	21	Yes	EM, BU, MS, SE
<i>Lomatium ambiguum</i>	Wyeth's biscuitroot	forb	14-Jul	10	Yes	EM

Continued

Table 4. (Cont'd)

Scientific name	Common name	Life form	Date of first mature seed	Duration of mature seed (weeks)	Observed microsite sensitivity (≥2 weeks)	Observed phase variability
<i>Lomatium triernatum</i>	nine-leaf lomatium	forb	9-Jun	12	Yes	FL, FR, MS
<i>Lupinus argenteus</i>	silvery lupine	forb	7-Jul	2	No	
<i>Lupinus sericeus</i>	silky lupine	forb	30-Jun	3	Yes	EM, BU, FR, FL, MS
<i>Melica bulbosa</i>	oniongrass	graminoid	14-Jul	10	No	
<i>Mertensia oblongifolia</i>	leafy bluebell	forb	14-Jul	3	N/A	
<i>Microseris nutans</i>	nodding microseris	forb	23-Jun	3	N/A	
<i>Microsteris gracilis</i>	pink microsteris	forb	9-Jun	5	Yes	EM, FL, SE
<i>Nemophila breviflora</i>	Great Basin blue-eyes	forb	30-Jun	3	N/A	
<i>Orthocarpus tenuifolius</i>	pink owl clover	forb	7-Jul	7	Yes	FL, FR, MS
<i>Packera cana</i>	woolly groundsel	forb	7-Jul	2	No	
<i>Pedicularis contorta</i>	coil-beaked lousewort	forb	7-Jul	3	N/A	
<i>Penstemon eriantherus</i>	fuzzy tongue penstemon	forb	14-Jul	10	N/A	
<i>Penstemon wilcoxii</i>	Wilcox's penstemon	forb	7-Jul	11	N/A	
<i>Perideridia gairdneri</i>	Gardner's yampah	forb	1-Sep	4	Yes	EM, BU
<i>Phacelia linearis</i>	thread-leaf phacelia	forb	14-Jul	2	Yes	FL, FR

Increased fecundity of exotic plants might be expected due to reduced competition for resources when the exotics emerge earlier and stay green later than native plants. Wolkovich and Cleland (2010) stated that the success of exotic species may be due to (1) temporally vacant niches, (2) being active earlier in the growing season, (3) having wider phenological niches, and/or (4) greater phenological plasticity. Our data support the first three of these statements, and cannot address the fourth as this is a single-year study. Pearson et al. (2012) found that in grasslands of western Montana, top invaders appear to exploit an empty temporal niche. Their study highlighted how exotics emerged earlier, but bolted and flowered later. Our study supports these results, and we also found that exotics had longer periods of total emergence and seed dispersal as well as differences in initiation of phenological phases. Knapp and Kühn (2012) also found that nonnative species in Germany were more likely to flower later than natives.

Research is emerging on quantifying the fitness importance of differing phenology of nonnatives. Godoy and Levine (2014) found that differing phenology conferred fitness advantage via the occupation of vacant niches, and Verdú and Traveset (2005) found early emergence of exotics conferred a fitness advantage. Future research with exotics in our ecoregion might address competition of exotics with natives of the most similar phenology.

Flowering Phenology of Native Plants

Native plants in our study area flowered between March and November. Native short-lived forbs, native perennial forbs, and native shrubs all showed a wide and similar variation in date of first flower and duration of flowering. Native graminoids had a shorter and less-varied flowering window. The temporal staggering of native flowering suggests that species occupy temporal niches. This supports a trait-based community assembly (Ackerly and Cornwell 2007).

When restoring a degraded system, reassembly of a complete flowering regime is important for ecosystem function, so

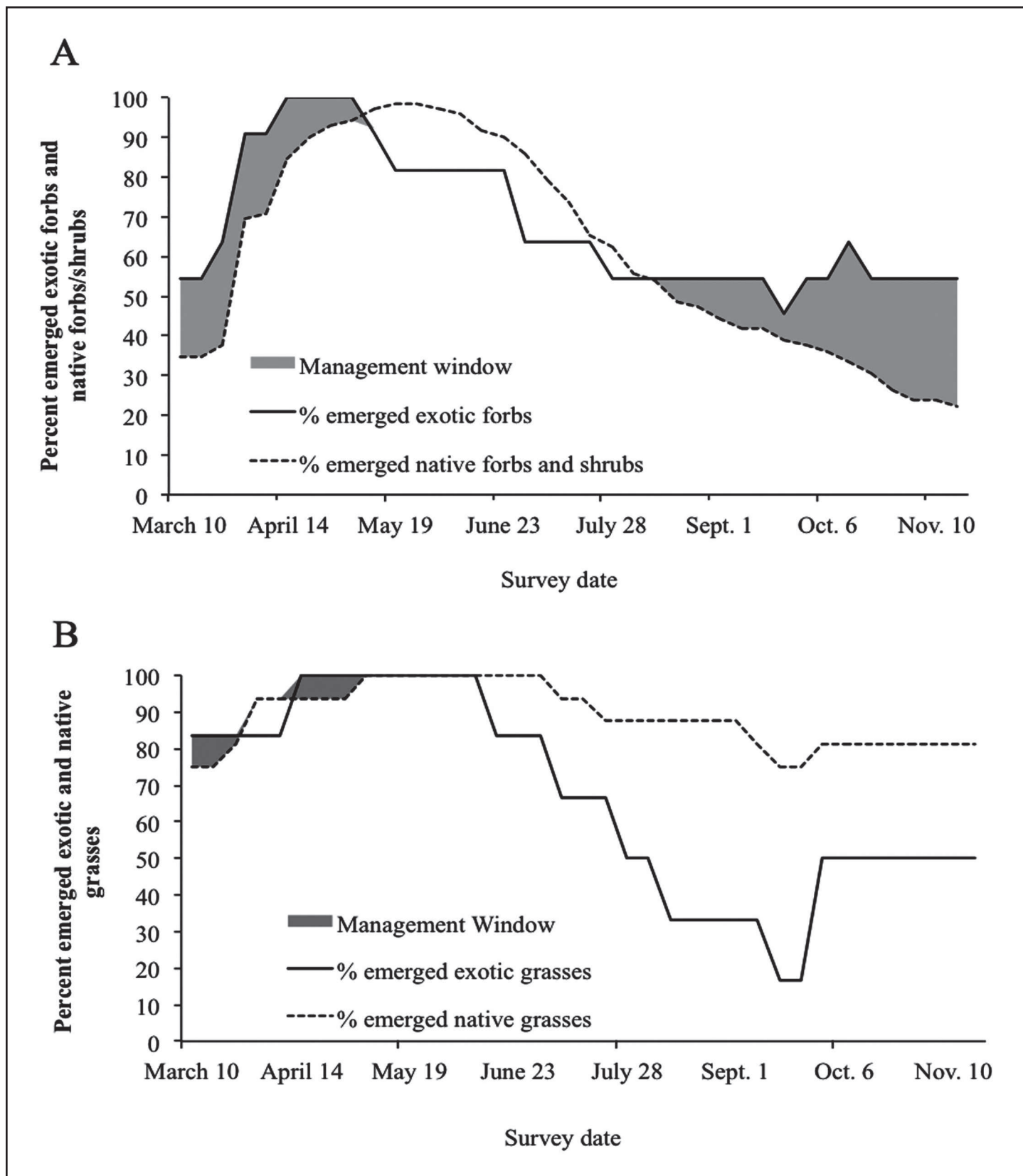


Figure 4. The percent of surveyed forbs and shrubs (A) and grasses (B) emerged at each weekly survey throughout the growing season. The dotted line represents the percent emerged of all surveyed native forbs and shrubs (A) and all surveyed native grasses (B). The solid line represents the percent emerged of all surveyed exotic forbs (A) and all surveyed exotic grasses (B). The “management window” (gray fill) is the period in time when the percent of exotic plants with vegetative growth is greater than the native plants, when control measures are presumed to harm the least number of native species.

attention should be given to capture a community assemblage with this wide span of temporal flowering.

Seed Collection and Genetic Materials

Restoration ecologists often research and consider conditions of origin when choosing seed (Krauss and He 2006; Barnes 2009). We wanted to know the extent of phenological variation across microsites to help us determine the relative importance of seed source for our collection efforts. We supply this information in table form in hope that it can be useful to others (Table 4).

Variables often used to consider seed source include the climatic conditions of rainfall, winter minimum temperatures, summer maximum temperatures, and soil types (Withrow-Robinson and Johnson 2006). However, through this filter our entire area of study would have nearly identical conditions (unpub. data). Microsites within the study area vary in total annual solar radiation, soil moisture, nutrient availability, and soil and air temperatures. Microsites can be important factors for plant establishment and success (Titus and Tsuyuzaki 2003; Dunwiddie and Martin 2016). Since plant phenology can vary with environmental cues, it follows that microsites, with their unique combination of light, temperature, and soil moisture, may also influence phenology. Species with different phenology across microsites may show phenotypic plasticity, or may have evolved genetically unique phenology at that microsite (Rathke and Lacey 1985; Richards et al. 2006). Genetically fixed phenology would retain the phenological patterns of origin, whereas species showing plasticity might adjust phenology to changing environmental cues. Without a common garden study for each species, we are unable to determine which populations have fixed genetic phenology. Consequently, species from the same elevation and climatic condition may prove unfit at another location because species may retain phenology when growing elsewhere (Godoy et al. 2009a). These species with microsite-dependent phenology and fixed genetic responses may require attention to seed source for restoration, even within a

local seed collection area.

Selection of seed source to maximize genetic fitness is often approached through provenances or elevation, and on such a scale that a nearby population is considered to have the same climatic conditions (Miller et al. 2011). A common assumption is that sites close together in space and elevation produce similarly adapted species (McKay et al. 2005). We found that for some species, phenology varied up to four weeks between proximal sites. For example, the shrub *Philadelphus lewisii* Pursh. (mock orange) flowered four weeks later at a site that differed by 85-m elevation and 1000-m distance. The late-flowering mock orange grew in the bottom of a steep V-shaped draw and received seasonal and daily shade, while the earlier flowering site was south facing and with many heat-holding rocks. The fact that so many native species show phenological variability suggests that, for this area, topography and insolation may create more specific environmental cues than ecoregion, soil type, geographic distance, or elevation.

In conclusion, we also note that seed collection sometimes occurs only in the late summer or fall. Our data indicate seed collection must occur throughout the entire growing season (starting as early as April) to capture the breadth of native diversity. Additionally, some species have available seed for only a week, so consideration of individual dispersal strategy is important for planning purposes.

Management Windows

Broadleaf herbicide applications in rangelands seek to benefit a desired plant community by suppressing a target exotic forb (Pearson and Ortega 2009). However, nontarget effects on desirable plants within the community and the release of exotic grasses can decrease the benefits of such applications (Ortega and Pearson 2011). For this reason, caution has been urged in the application of herbicides in wildland settings because the herbicide application does not always lead to native species recovery; rather, secondary invaders often occupy the treated area (Mason and French

2007; Rinella et al. 2009; Larson and Larson 2010; Kettenring and Adams 2011).

A range of environmental, physiological, and biochemical factors determine herbicide susceptibility (Monaco et al. 2002), but a prerequisite for foliar herbicide uptake is green leaf tissue (Wang and Liu 2007). One approach for exotic control is to apply herbicides while most natives are dormant (Cleland et al. 2013). However, this could harm susceptible remnant natives in competition with or evolving to compete with exotics in the same niche space (Fargione and Tilman 2005; Funk et al. 2008; Godoy et al. 2009b). To maximize competition with exotics, killing native species that share phenological patterns with the exotics should be avoided. Herbicide applications during the window when only a few natives are active could further reduce the native competition and allow for the establishment of exotics by increasing the empty temporal niche.

Our results indicate no clear window when natives are dormant while exotics are active in our study area. To illustrate, in the period when exotic forb presence was greatest and native forb presence was lowest, 19 of the 72 monitored native forb species were growing in at least one study site (Figure 4). Although this number is substantially lower than the number of species present during peak activity, if one were to apply herbicide in this period these 19 native species, which may be better adapted to tolerate exotic species, could be harmed. Native grass species remained emergent across the growing season; at the lowest points of emergence, 12 of 16 native grass study species were still present with green tissue in at least one study site. Even at one of the most degraded sites, with only 16% native plant cover, native grass and forb species were concurrently emergent with *C. stoebe* and *B. tectorum* from March through November. Removing native species from direct competition at that site could open niche space for exotics. Possibly, management windows could only be identified through careful observation by a manager near the time of herbicide application due to seasonal variability in differences between dormancy of native and exotic plants.

ACKNOWLEDGMENTS

We thank the MPG Ranch for support.

Rebecca A. Durham is a botanist at MPG Ranch. She holds a BA in Biology from Colby College and an MS in Botany from Oregon State University. Her thesis work investigated post-fire vegetation succession in the Bob Marshall Wilderness of Montana. Before coming to MPG, she gained experience performing restoration and vegetation research for the Forest Service, universities, consulting firms, and nonprofits. Her current research interests are plant phenology, plant community ecology, and biological soil crusts.

Daniel L. Mummey earned his BA in microbiology at Eastern Washington University, his MS in soil science at Washington State University, and his PhD in soil science and restoration ecology at the University of Wyoming. After graduating in 2004, he worked as an assistant research professor in the Department of Biological Sciences at the University of Montana, Missoula. Dan's research focuses on how soil structure, soil organisms, and plants determine plant community composition and ecosystem function. In Dan's current role at MPG Ranch, he develops and implements methods to establish healthy native plant communities in disturbed areas.

Lauren Shreading has a BA in biology from the University of Montana. She works in restoration and ecology research at MPG Ranch.

Philip W. Ramsey, PhD, is an ecologist and general manager of MPG Ranch. His research interests are in ecosystem processes and he has published on the influence of management practices on forest soils, factors allowing for the spread of invasive weeds in grasslands, and nutrient flow between rivers and floodplain forests. In addition to research, he oversees the management and operations of MPG Ranch, a 3800-hectare conservation property in western Montana.

LITERATURE CITED

Ackerly, D.D., and W.K. Cornwell. 2007. A

trait-based approach to community assembly: Partitioning of species trait values into within- and among-community components. *Ecology Letters* 10:135-145.

Ansquer, P., R. Al Haj Khaled, P. Cruz, J.P. Theau, O. Therond, and M. Duru. 2009. Characterizing and predicting plant phenology in species-rich grasslands. *Grass and Forage Science* 64:57-70.

Barnes, M. 2009. The effect of plant source location on restoration success: A reciprocal transplant experiment with winterfat (*Krascheninnikovia lanata*). PhD dissertation, University of New Mexico, Albuquerque.

Cadotte, M.W., and J. Lovett-Doust. 2002. Ecological and taxonomic differences between rare and common plants of southwestern Ontario. *Ecoscience* 9:397-406.

Cadotte, M.W., B.R. Murray, and J. Lovett-Doust. 2006. Evolutionary and ecological influences of plant invader success in the flora of Ontario. *Ecoscience* 13:388-395.

Cleland, E.E., L. Larios, and K.N. Suding. 2013. Strengthening invasion filters to reassemble native plant communities: Soil resources and phenological overlap. *Restoration Ecology* 21:390-398.

Dunwiddie, P.W., and R.A. Martin. 2016. Microsites matter: Improving the success of rare species reintroductions. *PLOS ONE* 11.3:e0150417.

Fargione, J., and D. Tilman. 2005. Niche differences in phenology and rooting depth promote coexistence with a dominant C₄ bunchgrass. *Oecologia* 143:598-606.

Franks, S.J., S. Sim, and A.E. Weis. 2007. Rapid evolution of flowering time by an annual plant in response to a climate fluctuation. *Proceedings of the National Academy of Sciences of the United States of America* 104:1278-1282.

Funk, J.L., E.E. Cleland, K.N. Suding, and E.S. Zavaleta. 2008. Restoration through reassembly: Plant traits and invasion resistance. *Trends in Ecology and Evolution* 23:695-703.

Ghersa, G.M., and J.S. Holt. 1995. Using phenology prediction in weed management: A review. *Weed Research* 35:461-470.

Godoy, O., and J.M. Levine. 2014. Phenology effects on invasion success: Insights from coupling field experiments to coexistence theory. *Ecology* 95:726-736.

Godoy, O., D.M. Richardson, F. Valladares, and P. Castro-Díez. 2009a. Flowering phenology of invasive alien plant species compared with native species in three Mediterranean-type ecosystems. *Annals of Botany* 103:485-495.

Godoy, O., P. Castro-Díez, F. Valladares, and M. Costa-Tenorio. 2009b. Different flowering

phenology of alien invasive species in Spain: Evidence for the use of an empty temporal niche? *Plant Biology* 11:803-811.

Jia, P., T. Bayaerta, X. Li, and G. Du. 2011. Relationships between flowering phenology and functional traits in eastern Tibet alpine meadow. *Arctic, Antarctic, and Alpine Research* 43:585-592.

Kettenring, K.M., and C.R. Adams. 2011. Lessons learned from invasive plant control experiments: A systematic review and meta-analysis. *Journal of Applied Ecology* 48:970-979.

Knapp, S., and I. Kühn. 2012. Origin matters: Widely distributed native and non-native species benefit from different functional traits. *Ecology Letters* 15:696-703.

Krauss, S.L., and T.H. He. 2006. Rapid genetic identification of local provenance seed collection zones for ecological restoration and biodiversity conservation. *Journal for Nature Conservation* 14:190-199.

Larigauderie, A., and J.H. Richards. 1994. Root proliferation characteristics of seven perennial arid-land grasses in nutrient-enriched microsites. *Oecologia* 99:102-111.

Larson, D.L., and J.L. Larson. 2010. Control of one invasive plant species allows exotic grasses to become dominant in northern Great Plains grasslands. *Biological Conservation* 143:1901-1910.

Lesica, P., and P.M. Kittelson. 2010. Precipitation and temperature are associated with advanced flowering phenology in a semi-arid grassland. *Journal of Arid Environments* 74:1013-1017.

Maron, J.L., S.C. Elmendorf, and M. Vilà. 2007. Contrasting plant physiological adaptation to climate in the native and introduced range of *Hypericum perforatum*. *Evolution* 61:1912-1924.

Mason, T.J., and K. French. 2007. Management regimes for a plant invader differentially impact resident communities. *Biological Conservation* 136:246-259.

McKay, J.K., C.E. Christian, S. Harrison, and K.J. Rice. 2005. "How local is local?" A review of practical and conceptual issues in the genetics of restoration. *Restoration Ecology* 13:432-440.

Miller, S.A., A. Bartow, M. Gisler, K. Ward, A.S. Young, and T.M. Kaye. 2011. Can an ecoregion serve as a seed transfer zone? Evidence from a common garden study with five native species. *Restoration Ecology* 19:268-276.

Monaco, T.J., S.C. Weller, and F.M. Ashton. 2002. *Weed Science: Principles and Practices*. 4th ed. John Wiley & Sons, New York.

Ortega, Y.K., and D.E. Pearson. 2011. Long-

- term effects of weed control with picloram along a gradient of spotted knapweed invasion. *Rangeland Ecology and Management* 64:67-77.
- Pearson, D.E., and Y. Ortega. 2009. Managing invasive plants in natural areas: Moving beyond weed control. Pp. 1-21 *in* R.V. Kingely, ed. *Weeds: Management, Economic Impacts and Biology*. Nova Publishers, New York.
- Pearson, D.E., Y.K. Ortega, and S.J. Sears. 2012. Darwin's naturalization hypothesis up-close: Intermountain grassland invaders differ morphologically and phenologically from native community dominants. *Biological Invasions* 14:901-913.
- Rathke, B., and E.P. Lacey. 1985. Phenological patterns of terrestrial plants. *Annual Review of Ecology, Evolution, and Systematics* 16:179-214.
- Rice, P.M., J.C. Toney, D.J. Bedunah, and C.E. Carlson. 1997. Plant community diversity and growth form responses to herbicide applications for control of *Centaurea maculosa*. *Journal of Applied Ecology* 34:1397-1412.
- Richards, C.L., O. Bossdorf, N.Z. Muth, J. Gurevitch, and M. Pigliucci. 2006. Jack of all trades, master of some? On the role of phenotypic plasticity in plant invasions. *Ecology Letters* 9:981-993.
- Rinella, M.J., B.D. Maxwell, P.K. Fay, T. Weaver, and R.L. Sheley. 2009. Control effort exacerbates invasive-species problem. *Ecological Applications* 19:155-162.
- Titus, T.H., and S. Tsuyuzaki. 2003. Distribution of plants in relation to microsites on recent volcanic substrates on Mount Koma, Hokkaido, Japan. *Ecological Research* 18:91-98.
- Verdú, M., and A. Traveset. 2005. Early emergence enhances plant fitness: A phylogenetically controlled meta-analysis. *Ecology* 86:1385-1394.
- Wainwright, C.E., and E.E. Cleland. 2013. Exotic species display greater germination plasticity and higher germination rates than native species across multiple cues. *Biological Invasions* 15:2253-2264.
- Wainwright, C.E., E.M. Wolkovich, and E.E. Cleland. 2012. Seasonal priority effects: Implications for invasion and restoration in a semi-arid system. *Journal of Applied Ecology* 49:234-241.
- Wang, C.J., and Z.Q. Liu. 2007. Foliar uptake of pesticides—present status and future challenge. *Pesticide Biochemistry and Physiology* 87:1-8.
- Withrow-Robinson, B., and R. Johnson. 2006. Selecting native plant materials for restoration projects: Ensuring local adaptation and maintaining genetic diversity. Publication EM 8885-E, Oregon State University Extension Service, Corvallis.
- Wolkovich, E.M., and E.E. Cleland. 2010. The phenology of plant invasions: A community ecology perspective. *Frontiers in Ecology and the Environment* 9:287-294.

Appendix. Flower duration for each study species by life form.

		Mar	Apr
		10 17 24 31	7 14 21 28
Perennial Forbs	Native		
	<i>Lomatium cous</i>	cous biscuitroot	
	<i>Ranunculus glaberrimus</i>	sagebrush buttercup	
	<i>Fritillaria pudica</i>	yellow bell	
	<i>Lomatium triternatum</i>	nine-leaf lomatium	
	<i>Dodecatheon conjugens</i>	Bonneville shootingstar	
	<i>Antennaria dimorpha</i>	low pussy-toes	
	<i>Erigeron compositus</i>	cutleaf daisy	
	<i>Hydrophyllum capitatum</i>	ball head waterleaf	
	<i>Lithophragma parviflorum</i>	smallflower woodland-star	
	<i>Claytonia lanceolata</i>	lanceleaf springbeauty	
	<i>Delphinium bicolor</i>	larkspur	
	<i>Viola nuttallii</i>	Nuttall's violet	
	<i>Dodecatheon pulchellum</i>	few-flowered shooting star	
	<i>Erythronium grandiflorum</i>	glacier lily	
	<i>Geum triflorum</i>	prairie smoke	
	<i>Lomatium ambiguum</i>	Wyeth's biscuitroot	
	<i>Lomatium macrocarpum</i>	big seed biscuitroot	
	<i>Balsamorhiza sagittata</i>	arrowleaf balsamroot	
	<i>Lithophragma glabrum</i>	smooth fringecup	
	<i>Antennaria rosea</i>	rosy pussy-toes	
	<i>Antennaria umbrinella</i>	umber pussytoes	
	<i>Saxifraga integrifolia</i>	wholeleaf saxifrage	
	<i>Senecio integerrimus</i>	lambstongue groundsel	
	<i>Zigadenus venenosus</i>	death camas	
	<i>Lithospermum ruderale</i>	field gromwell	
	<i>Arenaria congesta</i>	ballhead sandwort	
	<i>Lupinus sericeus</i>	silky lupine	
	<i>Triteleia grandiflora</i>	triteleia	
	<i>Astragalus inflexus</i>	hairy milkvetch	
	<i>Astragalus miser</i>	timber milkvetch	
	<i>Lupinus argenteus</i>	silvery lupine	
	<i>Penstemon eriantherus</i>	fuzzytongue penstemon	
	<i>Penstemon wilcoxii</i>	Wilcox's penstemon	
	<i>Agoseris glauca</i>	false-dandelion	
	<i>Arnica sororia</i>	twin arnica	
	<i>Eriogonum ovalifolium</i>	cushion buckwheat	
	<i>Silene oregana</i>	Oregon champion	
	<i>Apocynum androsaemifolium</i>	spreading dogbane	
	<i>Erigeron pumilus</i>	shaggy fleabane	
	<i>Lewisia rediviva</i>	bitterroot	
	<i>Microseris nutans</i>	nodding microseris	
	<i>Eriogonum umbellatum</i>	sulphur buckwheat	
	<i>Achillea millefolium</i>	western yarrow	
	<i>Crepis atriobarba</i>	slender hawksbeard	
	<i>Gaillardia aristata</i>	blanket flower	
	<i>Pedicularis contorta</i>	coil-beaked lousewort	

[illegible]

Appendix. (Cont'd)										
			Mar				Apr			
			10	17	24	31	7	14	21	28
Scientific name			Common name							
Short-lived forbs	Native	Potentilla glandulosa	sticky cinquefoil							
		Packera cana	woolly groundsel							
		Agastache urticifolia	giant horsemint							
		Allium cernuum	nodding onion							
		Crepis intermedia	gray hawksbeard							
		Heterotheca villosa	hairy golden aster							
		Campanula rotundifolia	scotch harebell							
		Erigeron speciosus	aspen fleabane							
		Potentilla gracilis	slender cinquefoil							
		Sedum stenopetalum	narrow-petaled stonecrop							
		Hieracium scouleri	Scouler's woollyweed							
		Perideridia gairdneri	Gardner's yampah							
		Solidago missouriensis	Missouri goldenrod							
		Exotic	Euphorbia esula	leafy spurge						
	Silene latifolia ssp. alba		bladder campion							
	Potentilla recta		sulphur cinquefoil							
	Centaurea stoebe		spotted knapweed							
	Native	Polemonium micranthum	littlebells							
		Microsteris gracilis	pink microsteris							
		Montia linearis	narrowleaved montia							
		Collinsia parviflora	blue-eyed Mary							
		Nemophila breviflora	Great Basin blue-eyes							
		Draba nemorosa	woods draba							
		Mertensia oblongifolia	leafy bluebell							
		Phacelia linearis	thread-leaf phacelia							
		Erigeron divergens	spreading fleabane							
		Collomia linearis	narrow-leaved collomia							
		Orthocarpus tenuifolius	pink owl clover							
Geranium viscosissimum		sticky geranium								
Clarkia pulchella		pink fairies								
Exotic		Draba verna	spring draba							
	Holosteum umbellatum	jagged chickweed								
	Erodium cicutarium	storksbill								
	Veronica verna	vernal speedwell								
	Alyssum alyssoides	yellow alyssum								
	Sisymbrium altissimum	tumble mustard								
	Camelina microcarpa	little-pod false flax								
	Tragopogon dubius	yellow salsify								
	Cynoglossum officinale	hound's tongue								
	Verbascum blattaria	moth mullein								
	Lactuca serriola	prickly lettuce								
	Carex geyeri	elk sedge								
	Carex petasata	Liddon's sedge								
	Poa secunda	Sandberg's bluegrass								
	Carex filifolia	threadleaf sedge								
	Koeleria macrantha	June grass								

Appendix. (Cont'd)																												
May				Jun					Jul				Aug				Sept					Oct				Nov		
5	12	19	26	2	9	16	23	30	7	14	21	28	4	11	18	25	1	8	15	22	29	6	13	20	27	3	10	17

			Mar 10 17 24 31	Apr 7 14 21 28
Perennial graminoids	Native	Scientific name	Common name	
	Exotic			
	Native	<i>Pseudoroegneria spicata</i>	bluebunch wheatgrass	
	Native	<i>Festuca campestris</i>	rough fescue	
	Native	<i>Festuca idahoensis</i>	Idaho fescue	
	Native	<i>Stipa occidentalis</i>	western needlegrass	
	Native	<i>Melica bulbosa</i>	oniongrass	
	Native	<i>Bromus carinatus</i>	mountain brome	
	Native	<i>Stipa comata</i>	needle and thread	
	Native	<i>Aristida purpurea</i>	three-awn	
	Native	<i>Danthonia unispicata</i>	onespike oatgrass	
	Native	<i>Bouteloua gracilis</i>	blue grama grass	
	Native	<i>Elymus glaucus</i>	blue wild rye	
	Exotic	<i>Poa bulbosa</i>	bulbous blue grass	
	Exotic	<i>Poa pratensis</i>	Kentucky bluegrass	
	Exotic	<i>Agropyron cristatum</i>	crested wheatgrass	
	Exotic	<i>Thinopyrum intermedium</i>	intermediate wheatgrass	
	Native	<i>Amelanchier alnifolia</i>	saskatoon	
	Native	<i>Acer glabrum</i>	Rocky Mountain maple	
	Native	<i>Purshia tridentata</i>	antelope bitterbrush	
	Native	<i>Prunus virginiana</i>	chokecherry	
	Native	<i>Crataegus douglasii</i>	black hawthorn	
	Native	<i>Physocarpus malvaceus</i>	ninebark	
	Native	<i>Philadelphus lewisii</i>	mock orange	
	Native	<i>Symphoricarpos albus</i>	snowberry	
	Native	<i>Chrysothamnus viscidiflorus</i>	green rabbitbrush	
	Native	<i>Ericameria nauseosa</i>	gray rabbitbrush	
	Native	<i>Artemisia dracuncululus</i>	wild tarragon	
	Native	<i>Artemisia frigida</i>	fringed sagebrush	
	Native	<i>Artemisia tridentata</i>	big sagebrush	
	Exotic	<i>Bromus tectorum</i>	cheatgrass	
	Exotic	<i>Bromus commutatus</i>	meadow brome-grass	
	Exotic	<i>Bromus japonicus</i>	Japanese brome	

Appendix. (Cont'd)																												
May				Jun				Jul				Aug				Sept				Oct				Nov				
5	12	19	26	2	9	16	23	30	7	14	21	28	4	11	18	25	1	8	15	22	29	6	13	20	27	3	10	17