

## **Pollen Limitation in the Alpine: A Meta-Analysis**

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# Pollen Limitation in the Alpine: a Meta-Analysis

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## Abstract

Pollen limitation (PL) is thought to be an important factor driving the evolution of floral traits in alpine plants. However, results of pollen supplementation experiments in alpine plants do not always show high levels of PL, and a general review of the importance of PL on alpine-plant reproduction is needed. We assessed to what extent alpine plants are pollen limited, and whether the intensity of PL differs between alpine and lowland species. Moreover, we assessed whether or not PL of alpine species depends on their reproductive system and, finally, whether the outcome of PL studies depends on whether supplemental-pollination was done at the whole-plant level or within a subset of available flowers. We performed both classical and phylogenetic meta-analysis. Our results show that alpine plants are pollen limited. However, PL did not differ significantly between alpine and lowland species. In the alpine, self-incompatible and self-compatible species show similar levels of PL. We did not find differences in PL between different manipulation levels. These results will help understand the real importance of PL of seed production in the alpine. We identify gaps in our knowledge of PL in the alpine that could serve to guide future directions of research in this field.

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## Introduction

The attraction of animal pollinators carrying an adequate quantity and quality of pollen is an essential part in the reproductive cycle of most angiosperm species. When plant reproductive success is reduced because of a shortage in the supply of pollen, it suffers from pollen limitation (PL), a widespread phenomenon among angiosperms (Burd, 1994; Larson and Barrett, 2000; Ashman et al., 2004). Several recent reviews have studied different aspects of PL. Among-species variation in PL depends on species' reproductive systems (PL self-incompatible species > PL self-compatible species), their habitats (PL in forested habitats > PL in open habitats), and distribution (PL in tropical areas > PL in temperate areas) (Larson and Barrett, 2000). Added to this, the extent of PL in a community is positively correlated with species richness (Vamosi et al., 2006), whereas the magnitude of PL in plant populations depends on both historical constraints and ecological factors (Knight et al., 2005). More recently, the theoretical models explaining the relationship between PL and abiotic resource availability have been improved, and some aspects of the ecological and evolutionary causes and consequences of PL have been clarified (Ashman et al., 2004). Moreover, reallocation processes have been revealed to be important in the results of PL experiments (Knight et al., 2006). Finally, the standard pollen-supplementation technique used to estimate PL has been recently criticized because it may not accurately estimate PL, confounding the effects of pollen quantity and quality limitation on reproduction (Aizen and Harder, 2007).

Intuitively, PL should be particularly strong in plant populations occurring in habitats where the availability of pollinators is low. One such habitat may be the arctic and alpine tundra, and indeed several earlier reviews of the ecology of arctic and alpine plants emphasize that their reproduction depends on pollination by insects (Kevan, 1972, 1973), thus it is strongly

challenged by a shortage of them (Müller, 1881; Mani, 1962; Billings and Mooney, 1968; Bliss, 1971; Kevan and Baker, 1983; Billings, 1987). As stated by Körner (1999) in his book on alpine plant ecology, "If one browses the more general alpine literature, one gets the impression that abundance and activity of pollinators decreases with elevation, so that alpine plants have a problem." Basically, the supposed strong PL in alpine and arctic species is attributed to severe weather conditions, such as low temperature and strong winds that restrict the flight activity of individual flower visitors, resulting in decreased flower visitation rates with altitude (Arroyo et al., 1985; Heinrich, 1993; McCall and Primack, 1992; Totland, 1994). The supposed strong PL in alpine and arctic species is believed to have wide-reaching evolutionary consequences. Low availability of pollinators has caused the evolution of relatively large flowers with particularly bright coloration (Billings and Mooney, 1968; Bliss, 1971) that supposedly enhance the ability of plants to attract the few pollinators available. Moreover, there is a higher frequency of plant species that use several different groups of pollinators (i.e. that are pollination generalists) in the alpine/arctic compared to elsewhere (Totland, 1993; Larson et al., 2001), and it could drive the evolution of pollinator-generalization in alpine species (Totland, 1993). With regard to sexual reproduction, very important in the alpine in terms of species numbers (Körner, 1999), and present even in the High Arctic flora (Kevan, 1972), the high level of inbreeding in alpine/arctic taxa has evolved because strong PL causes selection to favor selfing as a reproductive assurance mechanism (Crawford, 1989). The same causal pathway has been proposed to explain the high level of asexual reproduction in alpine species, including clonal reproduction and apomixis (Müller, 1881; Richards, 1997). Finally, with regard to pollination strategy, the high proportion of wind-pollinated or ambophilous (species using a mixture of wind- and animal-pollination) species in alpine/arctic habitats could be due to low pollinator service that selected for the evolution of

wind-pollination traits from animal-pollinated ancestors (Culley et al., 2002). Thus, the belief that pollinators are in limited supply in tundra habitats has had a prominent position in explaining the evolution of several floral traits that are believed to be typical of alpine and tundra plant species.

Nevertheless, during the last decade it has been proposed that the link between the frequency or quality of pollinator visitation and the resulting reproductive output may not be so obvious. Other factors than pollen availability, such as severe weather, low nutrient availability, and short growing season, also impact on the reproductive success of alpine plants and are suggested to constrain seed production to a level that prevents plants from taking advantage of a high deposition of pollen on their stigmas (Haig and Westoby, 1988; Zimmerman and Pyke, 1988; Ashman et al., 2004). The “Haig and Westoby equilibrium” (Ashman et al., 2004) between the resources allocated to pollinator attraction and seed provisioning predicts that there would not be an augmentation of seed set as a response to higher pollen loads. However, PL is common among flowering plants (Burd, 1994; Ashman et al., 2004; Knight et al., 2006), suggesting that departures from that equilibrium are common in nature. Burd (1995) explained it as a bet hedging strategy where flowers should be oversupplied with ovules relative to the average pollen load received, as an adaptation to stochasticity in pollen receipt. The alpine pollination environment is, as introduced above, highly unpredictable, so that plant populations in the alpine would depart from the Haig and Westoby equilibrium, and higher levels of PL could be expected in the alpine compared to lowland plant populations.

Added to this, to truly reveal PL in plants, it is necessary to supplement pollination of the majority of flowers on experimental individuals (hereafter plant-level experiments) (Zimmerman and Pyke, 1988). If the experiment is made only on a subset of the flowers (hereafter partial-level experiments), reallocation of resources from naturally pollinated to supplemented flowers on the same plant may lead to an overestimation of PL at the whole plant level (Knight et al., 2006), which is the relevant unit for evolutionary responses to pollen limitation. This issue may be particularly important in alpine species because their seed production may be strongly constrained by abiotic conditions that also could hamper resource reallocation processes.

The proposed causal pathway from severe weather conditions that result in a low pollinator visitation frequency, coupled by a low quality of visits (due to relatively inefficient pollinators), further causing a strong PL is apparently so obvious and convincing that few, if any, have critically opposed this idea. In this paper we use both classical and phylogenetically corrected meta-analysis to synthesize quantitatively experimental studies on PL in alpine plant species. In particular we ask: (1) Are alpine plants in general pollen limited? (2) Is the intensity of PL different in alpine compared to lowland species? (3) Does PL of alpine species depend on their reproductive system (self-compatible vs. self-incompatible)? (4) Does the outcome of PL studies depend on the manipulation level of the supplemental pollination experiment (plant-level vs. partial-level)?

## Methods

### SCOPE OF THE META-ANALYSIS AND DATA BASE BUILDING

We conducted a search of papers on the effect of cross-pollen addition on the reproductive output of arctic and alpine plants by means of hand-pollination experiments by using the ISI Web of Science (1945–2005). We used “poll\* limitation,” “supp\* poll\*,”

“hand poll\*,” “breeding system,” and “reproductive success” in combination with the terms “alpine” or “arctic” as search terms. We also used the lists of references within the papers on PL in alpine/arctic plants obtained through the ISI search and the paper collection of the authors to identify as many studies as possible that quantify PL in alpine/arctic species. We also included unpublished results if they fulfilled the requirements listed below.

We only included in the meta-analysis the articles fulfilling the following *a priori* requirements: (1) Hand-pollination experiments were accomplished using outcross pollen, and both control and treatment plants were open pollinated. We included one paper where the treated plants were bagged because the author explicitly noted that the treatment provided an excess of outcross pollen (Shykoff, 1988). We did not consider studies including emasculations of control or treatment plants. (2) The study was conducted under natural conditions. We did not include garden or glasshouse experiments because under such experimental conditions the pollinator assemblages would be modified. (3) The study included an estimate of female reproductive success to measure the plant response to the pollen-addition treatment. Among the different estimates, we included relative (seed-set, fruit-set, and seed/flowers ratio), and absolute (seeds per plant, seeds per flower, seeds per fruit) variables. When a single study reported different estimates of reproductive success for the same independent experiment we only included one of them in the analysis. We selected the variable most related to the success of seed production (mainly seed-set and analogous variables), in order to avoid bias attributable to the response variable used (Knight et al., 2006). (4) In studies where only indexes of reproductive success or PL (compounded from some of the above-mentioned variables) were reported as response variables of the hand-pollination experiments, we asked the authors for the original data used to calculate the indexes (Kasagi and Kudo, 2003).

As independent data points for the meta-analysis, we recorded data from experiments carried out at separate populations or at different years in the same population, as has been done in other meta-analyses on PL (Ashman et al., 2004; Knight et al., 2005) (see Appendix 4 [available free of charge at MetaPress: <http://instaar.metapress.com/content/120707> or at BioOne: <http://www.bioone.org/perlserv/?request=get-archive&issn=1523-0430>]).

From each study we recorded the means, the standard deviations (SD) and the sample sizes from both the pollen-addition treatment and the control. When only standard errors were available, we transformed them into SD. When the data were only available in figures, we used the ImageJ free software (<http://rsb.info.nih.gov/ij/>) to obtain the mean and SD directly from PDF files or after digitalizing the papers. When necessary, we asked the authors to clarify uncertainties or to provide missing data.

Among the experiments on PL obtained from the bibliographic search described above, we differentiated those in which authors noted that hand-pollinations had been done in more than or equal to 80% of the flowers in individual plants (plant-level) from those where only a small subset of flowers were manipulated (partial-level). With this data set, we tested the hypothesis that PL at the plant-level is lower than PL at the partial-level, presumably reflecting a resource limitation of reproduction via reallocation processes (Knight et al., 2006). Moreover, we obtained information on the breeding system (self-compatible vs. self-incompatible) of the species included in the meta-analysis according to the information given in the papers, to test the hypothesis that PL depended on the ability of plants to produce seeds by self-pollen and also potentially in the absence of pollinators. Finally, we tested the hypothesis that PL should be higher in alpine compared to lowland species. To do this, we compared Ashman et al.’s

(2004) data set (lowland sample) to our data set from the alpine (alpine sample). To enable this comparison, we modified Ashman et al.'s (2004) data set by removing the alpine species included in it and calculating Hedges'  $d$  (Gurevitch and Hedges, 2001) as the measure of effect sizes. In this comparison we used only our data from plant-level pollination experiments carried out in the alpine, because the Ashman et al. (2004) data set only included plant-level experiments.

## STATISTICAL ANALYSIS

We used Hedges'  $d$  (Gurevitch and Hedges, 2001) as the measure of effect size in our meta-analysis. In this study it calculates a standardized difference between the estimates of female reproductive success of the control and the pollen supplementation treatment. It has been used in a number of meta-analyses (e.g. Ashman et al., 2004; Maestre et al., 2005) and its properties are well known.

We used two different statistical approaches to test our hypotheses. First, we analyzed the data using classical meta-analysis, treating each data record in the data set as an independent sample. This approach does not account for the common ancestry of the species included in the analysis, resulting in a pseudoreplication bias if there is a phylogenetic signal in the data set (Gurevitch et al., 2001; Garland et al., 2005). To account for this, as a second approach, we also used a phylogenetically controlled meta-analysis (Verdú and Traveset, 2004, 2005).

We conducted the meta-analyses with MetaWin version 2.1.4 (Rosenberg et al., 2000). The  $Q$  statistic (Hedges and Olkin, 1985) was used to test the differences in the intensity of PL between experiments at the plant-level and those at the partial-level in alpine plants, and also to test whether there were differences in the levels of PL depending on the breeding system of the species. The  $Q$  statistic was also used to test whether there were differences in the intensity of PL between alpine and lowland plants. We used a randomization procedure (5000 iterations) for all significance tests.

We used random-effects models in all the meta-analyses. Such models assume that studies differ not only by sampling error, but also by a random component in effect sizes between studies, termed "pooled study variance" (Rosenberg et al., 2000). These models are more suitable than fixed-effects ones when analyzing ecological data. Fixed-effect models' assumption that all the observed variation is attributable to the sampling error is difficult to meet when using data from a broad range of studies (Maestre et al., 2005).

We used a resampling procedure (5000 iterations each) to generate bootstrap confidence intervals (95%) that are more conservative (wider) than parametric ones. Confidence intervals obtained by this method are recommended when the sample sizes are low and the assumptions of normality are not met (Adams et al., 1997; Gurevitch and Hedges, 1997).

We calculated Rosenthal's fail-safe numbers to test the presence of publication bias in the data sets. These numbers represent the number of non-significant, unpublished, or missing studies that would need to be added to a meta-analysis in order to change the results from significance to non-significance (Rosenberg et al., 2000). If the fail-safe number is larger than five times the sample size plus 10, it is safe to conclude that the results are robust with regard to publication bias (Verdú and Traveset, 2005).

We carried out phylogenetic meta-analysis as first outlined by Verdú and Traveset (2005). For this analysis the sample sizes are the number of species because the weighted average effect sizes are used. These weighted average effect sizes are single (cumulative) effect sizes per species, the weight being the inverse of the variance

(Rosenberg et al., 2000). By means of the PDAP software (T. Garland, Jr., P. E. Midford, J. A. Jones, A. W. Dickerman, R. Díaz-Uriarte, unpublished manuscript "PDAP manual") we examined whether a phylogenetic signal was present in the calculated effects sizes (see Verdú and Traveset, 2005; and the "PDAP manual" for further explanations). We did this by using phylogenetically independent contrasts and comparing the variance of the observed weighted average effect sizes with the variance of a set of permuted data. If the variance of the real data is lower than 95% of the variances of the permuted data, a phylogenetic signal exists. After this, we used the `lm.phylog` function in the `PHYLOGR` package for R written by R. Díaz-Uriarte and T. Garland for the Comprehensive R Archive Network (<http://cran.r-project.org>) to implement the phylogenetic information in the meta-analysis models. Using this function we carried out weighted generalized least squares in which the phylogenetic information is included by means of the phylogenetic trees. The significance tests are obtained by contrasting the observed values (weighted average effect sizes) against simulated distributions obtained by running 1000 permutations in the `PDSIMUL` module of the PDAP software (Garland et al., 1993). We included the inverse of the effect size variance as the vector of weights, as meta-analyses do. We used the on-line software utility `PhyloMatic` (Webb and Donoghue, 2004) to obtain the phylogenetic trees for the phylogenetic meta-analysis. The phylogenetic trees (Appendices 2 and 3), one for the alpine plants data set and one for the alpine-lowland plants data set, were constructed with Davies et al. (2004) phylogeny for angiosperms. We set the branch lengths to unity.

## Results

### THE ALPINE PLANTS DATA SET

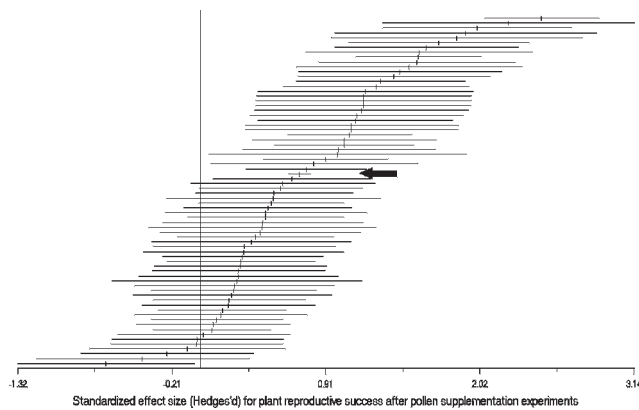
The data set comprised 18 studies including 24 species and 71 independent data records (Appendix 4). It contained 7 studies (studying 5 species and including 11 independent data records) where the hand-pollinations were carried out at the plant-level, and 11 studies (studying 19 species and including 60 independent data records) where the hand-pollinations were accomplished at the partial-level. In the data set, 13 species were self-incompatible ( $n = 40$ ) and 11 self-compatible ( $n = 31$ ). Forty-eight percent of the cases (and 50% of the species) were significantly pollen limited (Fig. 1). All the studies included in this data set were carried out in alpine habitats. We only found one paper studying PL in the Arctic (Philipp et al., 1996) and fulfilling all the *a priori* requirements of the bibliographic search. In this paper *Pedicularis lanata* is pollen limited ( $d = 0.67$ ,  $\text{var}(d) = 0.02$ ). We did not include it in further analysis, so that our analysis and discussion are focused on alpine plants.

The overall mean effect size of PL in alpine plants was 0.72 throughout with a bootstrap confidence interval that did not include zero (0.59–0.89), showing that PL in alpine plants is significantly higher than zero. Overall heterogeneity tests for this data set were non-significant (see below). The observed variance in the effect sizes (0.32) was lower than the 95% of the variances of the permuted data ( $p = 0.009$ ), showing the presence of a phylogenetic signal in the effect sizes.

### PLANT- VS. PARTIAL-LEVEL POLLEN LIMITATION

When comparing the hand-pollination manipulation levels, the total heterogeneity test was non-significant ( $Q_{\text{total}} = 67.33$ ,  $\text{df} = 69$ ,  $p = 0.60$ ), suggesting that the model explained the variance





**FIGURE 1.** Distribution of standardized effect sizes (Hedges'  $d$ , vertical bars) and sampling variances (horizontal bars) for 70 cases where pollen supplementation was conducted in alpine plants. Overall mean effect size of PL in alpine plants (arrow) was 0.74 with a bootstrap confidence interval of 0.59–0.89. Individual effect sizes with a sampling variance that do not overlap 0 are significant ( $p < 0.05$ ).

found in the data. Neither the Rosenthal's fail-safe number test ( $2135.5 > 360$ , fail-safe number  $>$  safe threshold), nor the weighted histogram (Fig. 2) denoted the presence of a publication bias. We did not find significant differences in PL between plant- ( $n = 11$ ) and partial-level pollen supplementation experiments ( $n = 60$ ) in PL ( $Q = 0.07$ ,  $df = 1$ ,  $p = 0.9$ ) (Fig. 3). This result is consistent when controlling for common ancestry (Coefficient =  $0.11$ ,  $F = -19.93$ ,  $p = 0.75$ ).

#### *POLLEN LIMITATION IN SELF-INCOMPATIBLE VS. SELF-COMPATIBLE ALPINE SPECIES*

The overall heterogeneity test was not significant ( $Q_{\text{total}} = 63.48$ ,  $df = 69$ ,  $p = 0.66$ ) when testing the effects of breeding system on PL, suggesting that the model explained the variance found in the data. Neither the Rosenthal's fail-safe number ( $2106.3 > 360$ ), nor the weighted histogram (Fig. 2) denote a publication bias. We did not find significant differences in PL between self-incompatible ( $n = 40$ ) and self-compatible species ( $n = 31$ ) ( $Q = 0.59$ ,  $df = 1$ ,  $p = 0.44$ ) (Fig. 3). The phylogenetically controlled meta-analysis did not show significant differences in PL between species with different breeding systems (Coefficient =  $0.18$ ,  $F = -19.91$ ,  $p = 0.73$ ).

#### *POLLEN LIMITATION IN ALPINE AND LOWLAND SPECIES*

This data set was formed by data from our subset of alpine plants (plant-level pollen supplementation, 7 studies, 5 species, and 11 independent data records) and the modified Ashman et al. (2004) data set (23 studies, 19 species, 79 independent data records). The result was a data set with 90 independent data points from 24 species (Appendix 5 [available free of charge at MetaPress: <http://instaar.metapress.com/content/120707> or at BioOne: <http://www.bioone.org/perlserv/?request=get-archive&issn=1523-0430>]).

The total heterogeneity test on this data set was non-significant ( $Q_{\text{total}} = 108.28$ ,  $df = 89$ ,  $p = 0.08$ ), so that our model explained the variability of the data. According to the Rosenthal's fail-safe number ( $1955.2 > 460$ ), and the weighted histogram (Fig. 2), there is no indication of a publication bias. The observed variance (0.45) was not lower than 95% of the variances of the

permuted data ( $p = 0.73$ ), so that there was no indication of a phylogenetical signal in the effect sizes of this data set.

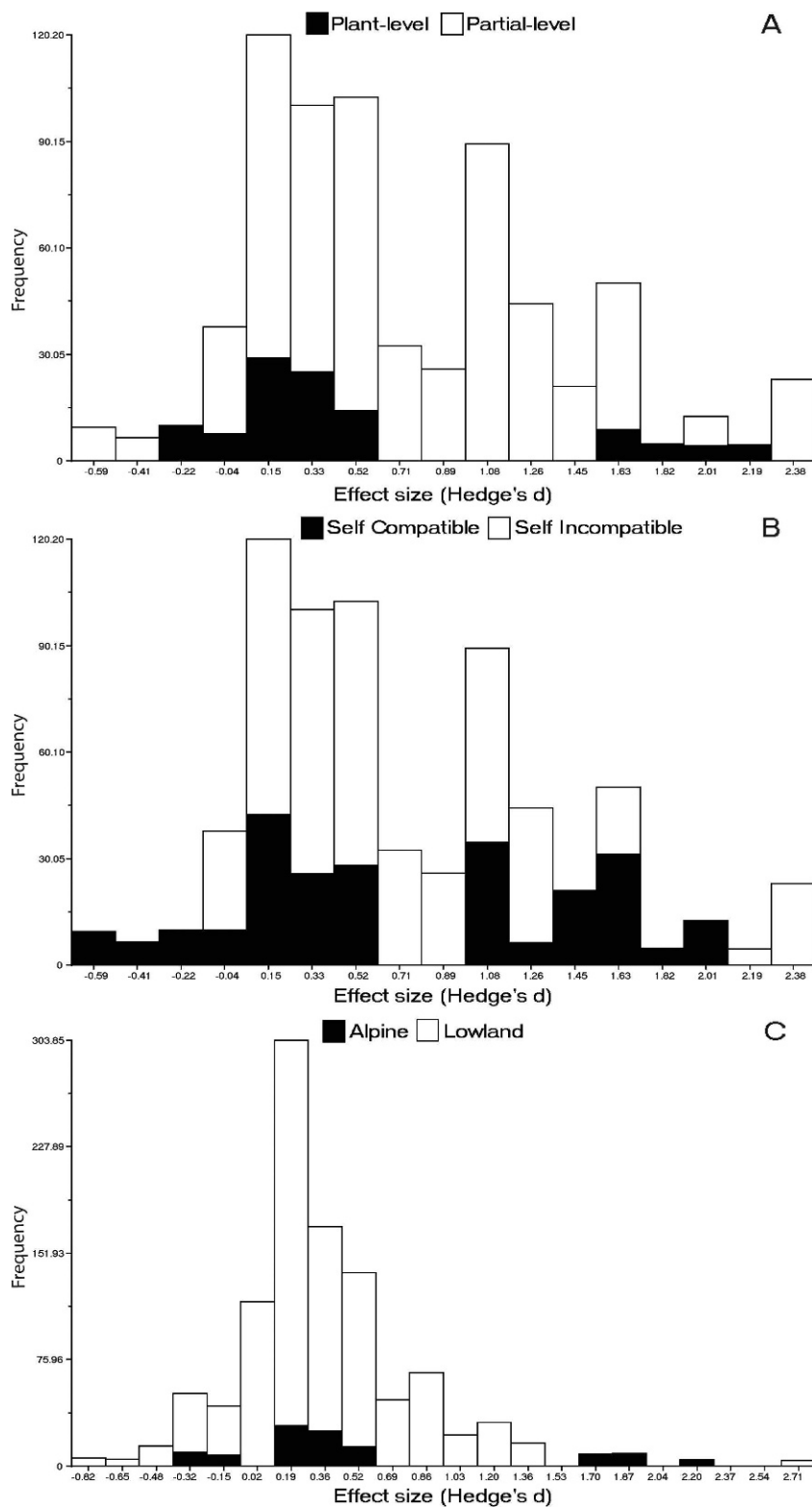
We did not find significant differences in PL between alpine ( $n = 11$ ) and lowland ( $n = 79$ ) plants ( $Q = 3.35$ ,  $df = 1$ ,  $p = 0.07$ ) (Fig. 3), a result consistent with the phylogenetically "informed" meta-analysis (Coefficient =  $-0.87$ ,  $F = 6.1$ ,  $p = 0.34$ ).

## **Discussion**

In general, 62–73% of plant species or populations have their female reproduction constrained by pollen availability (Burd, 1994; Ashman et al., 2004; Knight et al., 2006), showing that pollen limitation is a widespread phenomenon among the angiosperms. Our results showed lower percentages of prevalence of PL among alpine plants than in the data sets mentioned above: 48.05% of the cases (and 50% of the species) were significantly pollen limited. This percentage is still appreciable. We did not find significant differences in the levels of PL between alpine and lowland plants. The validity of this result is supported by the randomization method used to obtain it, the fail-safe numbers and the histogram. However, because of the low number of species examined for plant-level PL (5 alpine species for plant-level experiments, 11 independent data records) we have to be cautious about the conclusions drawn from this limited sample. These results do not agree with the presumption that PL should be particularly strong in alpine species primarily because severe weather conditions may constrain pollinator flower visitation activity.

We did not find differences in PL between plant- and partial-level pollen-supplementation experiments in alpine plants. It has been widely accepted that female reproductive success in flowering plants is limited both by pollen and resource availability (Haig and Westoby, 1988; Zimmerman and Pyke, 1988; Ashman et al., 2004). There are studies in the alpine reporting either PL (Galen, 1985; Miller et al., 1994) or resource limitation (Arft et al., 1999; Totland and Eide, 1999; Galen, 2000; Gaudeul and Till-Bottraud, 2004; Muñoz et al., 2005) as the main factors limiting female reproductive success. We expected PL to be higher in partial-level pollen-addition experiments than in plant-level ones, because in the partial-level reallocation of resources from other parts of the plant could increase female reproductive success of the hand-pollinated flowers at the expense of seed production in naturally pollinated flowers (Zimmerman and Pyke, 1988; Ashman et al., 2004; Knight et al., 2006). This difference between partial-level and plant-level pollen addition experiments, interpreted as driven by resource reallocation, has already been shown for an extensive data set of angiosperms (Knight et al., 2006). Future experimental studies in the alpine should focus on the relationships between PL and resource limitation, to test to what extent the seed production of alpine plants is limited by abiotic conditions, plants being unable to allocate extra resources to the supplementarily pollinated flowers. Apart from our evidence of the existence of PL in the alpine, there is other evidence supporting the idea that both PL and resources limit seed production in the alpine (Totland and Eide, 1999; Totland, 2001), the levels of PL in the alpine being lower than expected because resource limitation might constrain the use of extra-pollen loads by experimental plants.

We did not find differences in PL between studies using self-compatible and self-incompatible alpine species. This result contrasts with that of Larson and Barrett (2000) in an extensive data set across many habitats. They found lower levels of PL in self-compatible species than in self-incompatible ones. Our result suggests that self-compatibility in alpine species does not

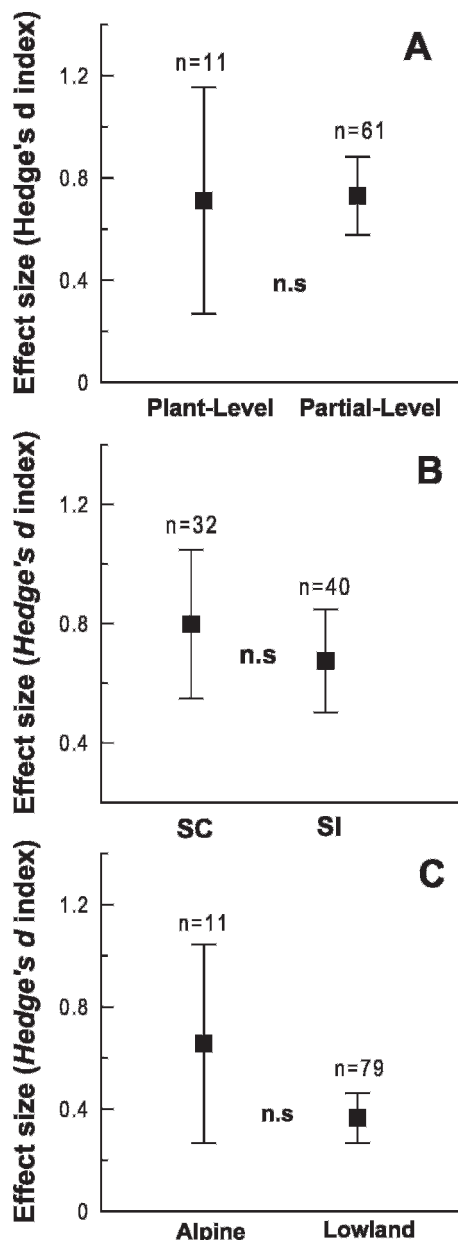


**FIGURE 2.** Weighted histograms of the effect sizes from pollen supplementation experiments in alpine plant species. (A) Plant-level (black), and partial-level experiments (white); (B) self-compatible (black), and self-incompatible alpine plant species (white); (C) alpine (black), and lowland plant species (white).

necessarily contribute to reduced PL. Some alpine species may have acquired self-compatibility before colonizing the alpine habitat, so that many of their traits (including their levels of PL) would not be a response to the alpine selective environment but rather an effect of phylogenetic inertia, as proposed by Arroyo et al. (2006). Furthermore, constraints from abiotic conditions could be invoked here as an additional explanation for the lack of difference in PL between self-compatible and self-incompatible species. Self-compatible species may potentially receive more

compatible pollen on their stigmas than self-incompatible ones, but certain levels of resource limitation might not let them take advantage of it.

Although plant species in the alpine are visited by insects even under the adverse conditions of nights (Kendall et al., 1981; Kevan and Kendall, 1997), pollinator visitation rates to plants in alpine habitats appear to be relatively low compared to plants at lower altitudes (Arroyo et al., 1982, 1985; Inouye and Pyke, 1988; Totland, 1993), possibly because of environmental constraints on



**FIGURE 3.** Mean effect sizes of PL and 95% bootstrap confidence intervals. (A) Plant-level ( $n = 11$ , 5 species) and partial-level ( $n = 60$ , 19 species); (B) SC (self-compatible) ( $n = 31$ , 11 species) and SI (self-incompatible) alpine species ( $n = 39$ , 13 species); (C) alpine ( $n = 11$ ) and lowland plants ( $n = 79$ ). Squares represent mean effect sizes of PL and vertical lines 95% bootstrap confidence intervals.  $n$ : the sample size, corresponding to the number of independent data records; n.s.: no significant differences between groups.

insect activity in alpine ecosystems (McCall and Primack, 1992; Stenström and Bergman, 1998; Totland, 1994). However, even if flowers of alpine plants are less visited by pollinators, they may benefit from a higher efficiency of the pollinators compared to those at lower altitudes, primarily through the higher abundance of efficient bumblebees (Arroyo et al., 1985; Galen and Stanton, 1989; Bingham and Orthner, 1999; but see Totland, 1993). Moreover, alpine plants may have an extended flower longevity compared to lowland plants, (Arroyo et al., 1985; Bingham and Orthner, 1999), which may compensate for a lower visitation frequency in terms of their reproductive success. These factors,

related to the pollination environment, combined with abiotic constraints on reproduction, may serve to explain the levels of PL in alpine species.

There is clearly a need for more studies on PL in alpine plants before we can assess with certainty the general importance of pollinators for seed production in alpine species. Moreover, PL studies conducted on alpine species on the plant-level are scarce, as well as data on pollinator visitation rates. We suggest that a fruitful approach to the study of PL in the alpine would be to conduct replicated supplementary pollination experiments along altitude gradients, supplemented by measurements of pollinator visitation rates and pollen deposition along the same gradient. Those future experiments should take into account Aizen and Harder's (2007) suggestions to avoid under and overestimation of PL, and also distinguish between pollen quantity and quality limitation of seed production. Furthermore, there is a need to reduce the "geographical bias" in studies of PL, because as far as we know, there is a lack of studies in the Arctic and on tropical mountains. Nevertheless, despite both methodological and geographical limitation of PL studies on alpine plants, the results of these studies, as revealed by our meta-analysis, should serve to stimulate a new critical look at the evolution of pollination and reproductive traits in alpine plants.

## Acknowledgments

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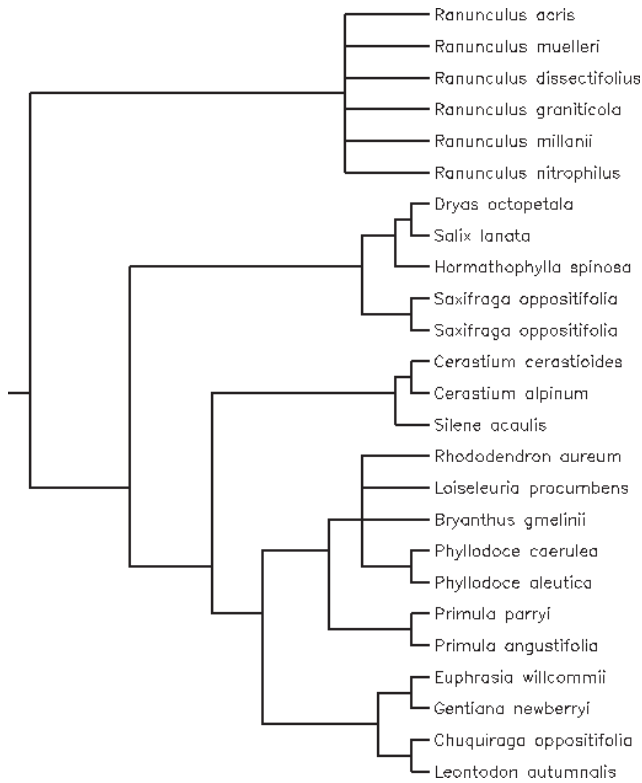
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## APPENDIX 2

Phylogenetic tree including the species of the alpine plants data set obtained using the online software utility Phylomatic (Webb and Donoghue, 2004) and constructed with the Davies et al. (2004) phylogeny for angiosperms. We set the branch lengths to unity.



## APPENDIX 3

Phylogenetic tree including the species of the alpine-lowland plants data set obtained using the online software utility Phylomatic (Webb and Donoghue, 2004) and constructed with the Davies et al. (2004) phylogeny for angiosperms. We set the branch lengths to unity.

