

Reframing Pastoral Practices of Bofedal Management to Increase the Resilience of Andean Water Towers

Authors: Yager, Karina, Prieto, Manuel, and Meneses, Rosa Isela

Source: Mountain Research and Development, 41(4)

Published By: International Mountain Society

URL: https://doi.org/10.1659/MRD-JOURNAL-D-21-00011.1

BioOne Complete (complete.BioOne.org) is a full-text database of 200 subscribed and open-access titles in the biological, ecological, and environmental sciences published by nonprofit societies, associations, museums, institutions, and presses.

Your use of this PDF, the BioOne Complete website, and all posted and associated content indicates your acceptance of BioOne's Terms of Use, available at www.bioone.org/terms-of-use.

Usage of BioOne Complete content is strictly limited to personal, educational, and non - commercial use. Commercial inquiries or rights and permissions requests should be directed to the individual publisher as copyright holder.

BioOne sees sustainable scholarly publishing as an inherently collaborative enterprise connecting authors, nonprofit publishers, academic institutions, research libraries, and research funders in the common goal of maximizing access to critical research.

Reframing Pastoral Practices of *Bofedal*Management to Increase the Resilience of Andean Water Towers

Karina Yager¹*, Manuel Prieto², and Rosa Isela Meneses³

- * Corresponding author: karina.yager@stonybrook.edu
- ¹ School of Marine and Atmospheric Sciences, Stony Brook University, Endeavour 145, Stony Brook, New York 11790, USA
- Departamento de Ciencias Históricas y Geográficas, Facultad de Educación y Humanidades, Universidad de Tarapacá, Avenida 18 de Septiembre N°2222, Casilla 7-D, Arica, Chile
- ³ Instituto de Arqueología y Antropología, Universidad Católica del Norte, Le Paige 380, San Pedro de Atacama, Chile

© 2021 Yager et al. This open access article is licensed under a Creative Commons Attribution 4.0 International License (http://creativecommons.org/licenses/by/4.0/). Please credit the authors and the full source.

Across the Andes, a critical challenge for mountain socioecological systems is securing water for future generations. Pastoral communities are especially vulnerable because their livelihood practices are often unseen or perceived as a threat to natural resource conservation. In addition to the challenges of climate change, socioeconomic and political processes complicate the drivers of pasture degradation and sustainable water management. Often overlooked systems in assessments of Andean water towers are bofedales (high-altitude peat wetlands), which are critical to supporting mountain pastoral livelihoods. While "natural" azonal mountain peatland and humid meadow development occurs across the Andes, we posit that bofedales are sociohydrological systems created through pastoral management practices over generations. Drawing on the results of applied research on bofedales across the Andes and a literature review of published papers, we present a conceptual reframing of bofedal typologies and change analysis, which prioritizes the role of pastoralists in interdisciplinary research and comparative

assessments of land-use and land-cover change in Andean highland regions. We identified key socioecological challenges to sustainable bofedal management, related to herder decision-making and articulated within broader socioeconomic processes. Reframing bofedales as sociohydrological constructs permits the identification of actionable knowledge and the support of water conservation practices applied by pastoralists across Andean water tower regions. If Andean pastoralists are recognized as stewards of sociohydrological systems that are critical to water towers, rather than perceived as threats to natural resources, bofedal conservation planning may be prioritized and locally supported.

Keywords: pastoralism; Andes; bofedales; water towers; cultural landscape; irrigation; mountain sustainability; sociohydrology; alpine peatlands.

Received: 19 February 2021 Accepted: 22 October 2021

Introduction

Across the Andes, threats of water scarcity are recognized as diminishing water supplies compounded by the impacts of climate change and increasing water demand by human populations and extraction industries, including urbanization, agriculture, hydropower, and mining (Orlove et al 2008; Urrutia and Vuille 2009; Bury et al 2013). Due to the climatic, orographic, and ecosystem characteristics, the tropical Andean mountains are recognized as "water towers" that store and regulate hydrological assets that are critical for highland ecosystems and downstream populations (Immerzeel et al 2020). Yet, Andean water towers are among the most vulnerable features to the impacts of climate change, marked by rapid glacier loss, diminishing discharge, and increased surface temperature, which increase sociohydrological risks (Mark et al 2017). As such, there is an urgent need to identify actionable pathways to secure and extend the sustainability of regional water supplies provided by Andean water towers (McDowell et al 2019).

In the high-elevation regions of the Andean water towers, pastoral communities have managed local hydrology and

mountain ecosystem resources for millennia (Capriles and Tripcevich 2016). Ethno-historical and archaeological studies have identified sophisticated sociopolitical organization regarding irrigation practices among Andean societies (Mitchell and Guillet 1994), yet few traditional practices remain intact. Many of these practices are vulnerable to loss, including pastoral management practices that remain unseen, undervalued, or even negatively perceived (Lane 2006; Verzijl and Quispe 2013). As Andean landscapes continue to rapidly transform, some irreversibly, the identification and continuation of local indigenous knowledge and practices that increase the resilience of mountain sociohydrological systems and support pastoral livelihoods are urgent pursuits (Gilles et al 2013; Valdivia et al 2013).

The observed climate data and model scenarios for the Andes show accelerating deglaciation, increases in surface temperature, and significant alterations in the seasonality and intensity of precipitation and extreme events (Buytaert et al 2010, 2011; Rabatel et al 2013). Across the high-elevation regions of the Andes, glaciers constitute a critical hydrological asset through the storage of water in snow and

ice and the seasonal release of outflow to support mountain ecosystems. In addition to monitoring physical change (eg surface area and volume) in hydrological systems (eg glaciers, lakes, and rivers), it is critical to conserve the regulating function of mountain ecosystems for sustaining water systems. Such ecosystems include peatlands (ie bofedales) and wetlands, meadows and grasslands (ie puna and paramo), and native forests (ie Polylepis spp). Due to the hydrological regulation and provision of bofedales (Segnini et al 2013), their sustainability in mountain regions under current climate change is widely recognized as a research priority (Bury et al 2013; Otto and Gibbons 2017; Polk et al 2017). In addition to the multiple socioecosystem services provided by bofedales, they play a significant role in carbon cycling, having some of the highest rates of sequestration of mountain landcover classes (Chimner and Karberg 2008; Buytaert et al 2011; Hribljan et al 2015).

Andean pastoral communities have managed *bofedales* since pre-Hispanic times (Flores-Ochoa 1977; Erickson 2000; Lane 2014; Capriles and Tripcevich 2016). In highland regions where pastoralism persists, indigenous communities (eg Aymara, Quechua, Colla, and Atacameños) depend upon and actively manage *bofedales* as a critical source of perennial green forage and water for herds of llama (*Lama glama*) and alpaca (*Vicugna pacos*), as well as nonnative cattle, sheep, horses, and goats (Browman 1989; Baied and Wheeler 1993; Villagrán and Castro 1997; Postigo et al 2008).

Interwoven with the impacts of climate change, local decision-making on pastoral management is influenced by broader socioeconomic systems and often results in outmigration, labor shortages, environmental contamination, and conflicts regarding land tenure, water access, and natural resource use (Coppock et al 2017; Figueroa-Armijos and Valdivia 2017). Once considered isolated mountain systems (Flannery et al 1989), Andean pastoral communities and mountain socioecological systems are now closely connected to production needs for globalized markets, for example, the luxury textile industry (eg vicuña and alpaca wool), international export of subsistence crops (eg quinoa), water extraction for mining (eg metals and lithium), and dam development for hydropower and urban populations. Nonproximate (ie regional or global) processes can also trigger unsustainable pastoral practices, such as increased grazing pressures, reduced pasture rotation, loss of traditional knowledge, and abandonment of communal irrigation practices. As a result, Andean pastoral management objectives, social relations, and risk management strategies have significantly changed in recent years, impacting land use, land-cover change (Coppock et al 2017), and the sustainability of bofedal systems (Yager et al 2019). In severe circumstances, pastoral communities experience displacement, shortages of social services, threats to economic welfare, exposure to environmental toxins, and increasing ethnic and gender inequalities.

Across the Andes, conflicts regarding water supply, demand, and allocation are palpable among pastoral communities. Viewing water management challenges through a transdisciplinary lens allows the identification, prioritization, and analysis of generalizable patterns that can facilitate a better understanding of the social, institutional, and economic activities of sociohydrological systems leading to sustainable hydrological outcomes (Brelsford et al 2020).

Water management decisions are cumulative, occurring at multiple institutional scales, through which infrastructure and practices are socially and politically negotiated to affect the control of water, including quantity, quality, allocation, and flow, the sum of which produces a sociohydrological system. Similarly, we posit that *bofedales* are key sociohydrological systems that are critical to the long-term sustainability of Andean water towers. As such, an appropriate transdisciplinary research framework is needed to evaluate the sociohydrological functions of *bofedales*, their management by mountain communities, and the processes that support sustainable outcomes.

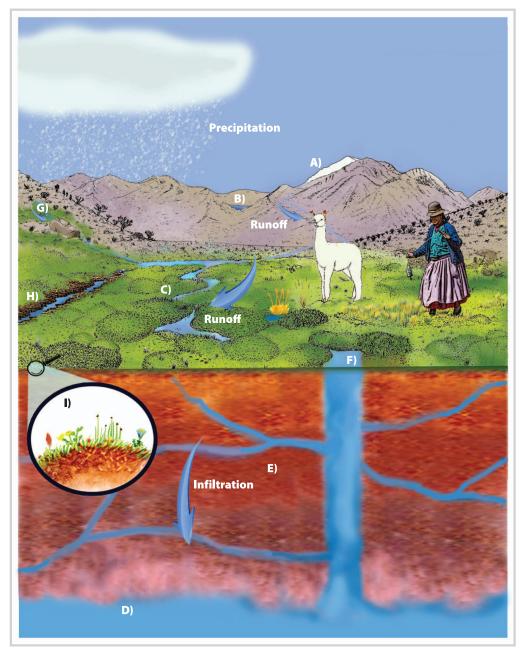
Pastoralists manage mosaics of mountain ecosystems, including bofedales, highland grasslands (puna and paramo), and native forests, which are all important for Andean natural resource use and conservation. A transdisciplinary research agenda producing translational knowledge on the integration of socioecological systems and practices is therefore necessary (Mathez-Stiefel et al 2017). We posit that reframing bofedales as sociohydrological systems is critical to secure sustainable water and forage resources for herding communities. Denaturalizing bofedales, in part, is necessary to understanding political and socioeconomic drivers of landuse and land-cover change, and to identifying the nested social institutions at multiple scales that influence the sustainability of bofedales. It is also necessary to include local stakeholders in water governance and conservation planning aimed at increasing the resilience of Andean water towers. In particular, a transdisciplinary, multiscale approach is needed to understand changing pastoral management decisionmaking at the microscale (eg within a herding parcel or sayaña), and in relation to local (eg community-level) and regional climate, political, and socioeconomic processes (eg water-extraction activities).

Background on bofedales

Andean water towers are best known for their towering glaciers, alpine lakes, and rivers that provide critical water supplies for mountain and downstream populations. Less visible but essential features are the mountain aquifers and natural springs, which are recharged by seasonal precipitation and snow events. Infiltration and replenishment of hydrological channels and interconnected systems are frequently dependent upon stable soils and adequate vegetation cover, including bofedales, which are often nestled in glacier valleys and at the bases of mountain flanks. Applying a broad natural science typology, bofedales are concentrated mosaics of compact cushion plants with low-growing plant assemblages associated with a network of mountain surface hydrological systems, including streams, water holes, and springs. Bofedales contribute to increasing water infiltration, recharge of aquifers, and slowing seasonal water runoff. As such, bofedales provide critical water regulation and storage for mountain socioecological systems and constitute key hydrological components of Andean water towers (Figure 1).

Commonly applied natural science descriptors for *bofedales* include mountain fens, mires, bogs, humid meadows, and alpine wetlands; importantly, each of these typologies differs in their distinct chemical, biogeographic, biotic, and hydrological properties (Lindsay 2018). While *bofedales* are

Figure 1 Schematic illustration of a bofedal and sociohydrological components in an Andean water tower, including precipitation, runoff, and infiltration contributing to water flow and recharge in a bofedal, and key landscape features: (A) glacier, (B) high-elevation lake, (C) bofedal, (D) aquifer, (E) organic material, (F) water hole (ojo de agua), (G) water spring, (H) irrigation canal, and (I) macroscale view of bofedal plants.



found in high-elevation alpine watersheds across the tropical Andes, there are generalizable differences from a natural science perspective in terms of their biogeographic distribution and characteristic vegetation composition.

In the arid to semiarid high-elevation regions of the tropical Central Andes, *bofedales* are azonal ecological communities occurring in the high *puna* region (at elevations of approximately 3500 to 5000 m above sea level [masl]) across Bolivia, Peru, northern Argentina, and Chile (Ruthsatz 2012; Meneses et al 2019). *Bofedales* are azonal in that they differ from the drier mountain land-cover classes due to year-round plant growth, organic soils, and provision of a constant source of water. They are otherwise termed a mountain peatland system. Through a botanical lens,

bofedales in the central and southern puna regions are composed of mosaics of compact cushion plants, dominated by vascular plants from the Juncaceae family (eg Oxychloe andina, Distichia spp, Patosia clandestina), and commonly associated with the presence of Cyperaceae (eg Phylloscirpus, Zameioscirpus), Plantaginaceae (Plantago tubulosa), and Gramineae (eg Deyeuxia spp, Poa spp) (Ruthsatz 1993, 2012; Luebert and Pliscoff 2006; Meneses et al 2015, 2019). In the northern puna–paramo transition of the tropical Andes, including Peru, Ecuador, and Colombia, the study of bofedales has been applied to describe both peat-accumulating systems (Chimner and Karberg 2008) and more broadly to mountain wetland systems (Polk et al 2017; Chimner et al 2019). The bofedales of the paramo and northern humid puna regions of

Ecuador and northern Peru have a higher cover of mosses in their vegetation associations (Cooper et al 2010), while the southern *puna* regions are associated with dry grasses (Cooper et al 2010, 2015; Meneses et al 2019). *Bofedales* vary in size according to their geomorphological setting and hydrological conditions (Squeo et al 2006). The water that replenishes *bofedal* systems may include contributions from surface hydrology (eg lakes and streams), springs, precipitation (including surface runoff), and glacier outflow (Polk et al 2017; Cooper et al 2019). In pastoralist communities, herein emphasized, the water is sustained by the construction of canals and waterways to redirect hydrological flow and infiltration.

From a broader social science perspective, the term "bofedales" is a local, vernacular description of high-elevation irrigated pastures used by indigenous Andean herding communities. Aymara and Quechua names for bofedales include ok'os or ughu, and the dominant plant species are called cachu (Distichia muscoides), kunkuna (Distichia muscoides), and k'uli urcu (Oxychloe andina), in addition to many associated plants (eg tiña, chinka, llachhu, kachu, waricha, porqu'e) (Palacios-Ríos 1977; Canales and Tapia Núñez 1987; Villagrán and Castro 1997). Other vernacular terms used to describe humid meadows include cienega, vega, or humedal, which are more prevalent in midelevation Andean regions (between 3000 and 4000 masl). The latter terms differ from bofedales in describing areas with limited forage value that do not accumulate peat and are more often unmanaged. This is in part due to their limited occurrence in agropastoral zones, where labor investment is typically more focused on crop production. Thus, there are important biogeographical, botanical, and sociocultural differences among commonly applied bofedal typologies.

Conceptual framing of bofedales

Given the broad range of terms and characteristics, researchers must consider the use and application of bofedal typologies, which may directly impact land-use decisions, policy priorities, and conservation planning. The conceptual framing of bofedales across a spectrum from "natural" to "anthropogenic" varies among stakeholders, policymakers, and scientists and thereby confounds the multiplicity of drivers of environmental change, as well as the perceived threats to the natural resources of Andean water towers. Institutional perceptions of land-use and land-cover change—in particular, drivers of degradation, erosion, and threats to biodiversity—have often been based on a dichotomous human-nature view, resulting in misidentification of indigenous practices and blame of local land-use managers as destructive agents (Fairhead and Leach 1996; Dove 2004, 2006). Such erroneous conceptual framing by state agencies can result in decadal- to century-long systems of sociopolitical oppression of local land users that are closely tied to natural resource management policies (Blaikie 1985; Dove and Kammen 2015).

In order to evaluate the conceptual framing of "bofedales" in current scientific literature, we conducted a bibliographic analysis of published research on bofedales in Chile, Argentina, Peru, and Bolivia. The study included keyword searches of titles and abstracts using reference databases (including the Web of Science Core Collection and SciELO)

Citation Index) to identify published journal articles, book chapters, and conference papers focused on bofedales. Our search identified 119 publications, published between 1977 and 2018, which included bofedal terminology and common synonyms. The publications were further analyzed using a coding process based on various bofedal characteristics (including location, methodologies applied, discipline, role of herder management, and vegetation types). From this study, a quantitative analysis of the scope of bibliographic references was carried out to further consider the breadth of applied disciplinary views and research approaches on these Andean ecosystems (see White-Nockleby et al 2021). Many of the fundamental natural science publications on bofedales are appropriately dedicated to localized botanical study of these systems (eg Ruthsatz 1993, 2012), and subregional peatland distribution and change analysis (eg Izquierdo et al 2015; Dangles et al 2017; Chimner et al 2019). Researchers from the natural sciences may consider the primary drivers of bofedal change to be bioclimatic and environmental, for example, geological, physiographical, morphological, climatological, or biological characteristics (Earle et al 2003; Squeo et al 2006; Dangles et al 2017). From an alternative perspective, researchers may confound all Andean wetlands as bofedales or erroneously conceive them as ahistorical and ecological settings of social relations (Hartman 1996; Gandarillas et al 2016; Struelens et al 2017). Both approaches reproduce a binary relationship between nature and culture; bofedales are not a strictly uniform biogeographical entity nor are human actors (ie herders) merely passive users of these ecosystems.

Over more than a decade, we have conducted applied research on bofedales with local communities in Peru, Bolivia, and Chile (eg Meneses et al 2015, 2019; Prieto 2015; Yager 2015; Prieto et al 2019; Yager et al 2019), including vegetation studies, institutional analysis, ethnographic research, geospatial analysis, hydrological studies, archaeological studies, and archival research. While publications may be appropriately focused on disciplinespecific topics, for example, botany (eg Meneses et al 2019) or hydrology (eg Cooper et al 2015), we recognize the value in identifying the linkages, drivers, and networks between and across systems. Interdisciplinary research on bofedales increasingly recognizes the critical role of local stakeholders and bofedales in mountain sustainability (eg Postigo et al 2008; Valdivia and Yager 2018). Furthermore, local testimonies from Andean communities recognize that water access and irrigation are critical to the sustainability of bofedales (Yager et al 2019). The impediment of a binary or bounded research approach may result in overlooking the imperative role that pastoralists have held over multiple generations in sustaining bofedales as sociohydrological systems. Bofedales are neither strictly "natural" nor "pristine" (Denevan 2001), but rather part of the built cultural landscape, or "landscape capital" (Erickson 2000), resulting from hundreds of years of management actions, embedded with indigenous knowledge of ecosystem linkages among climate, water, and vegetation. In this applied perspective, bofedales constitute fusion landscape entities and socioenvironmental encyclopedias of the ways in which cultural practices, socioeconomic institutions, and power relations impacting pastoralists have been negotiated over time under changing environmental, political, and socioeconomic influences.

Reframing bofedal management

Bofedales are mountain sociohydrological systems that are vital to mountain sustainability for highland and downstream communities. The built water infrastructure increases water storage, regulates outflow, slows erosion during extreme weather events, and is a critical source of water availability throughout the year, most importantly during the extended dry season or drought years (Garcia and Otto 2015). We posit that while zonal mountain peatland development occurs across the Andes, the alpine systems become a bofedal through intentional pastoral practices occurring over multiple generations. Bofedal condition reflects not only shifts in climate and environmental factors, but it also signifies change in dominant socioeconomic relationships, such as cultural knowledge, pastoral identity, community cohesion, and indigenous livelihood practices in relation to broader political and socioeconomic processes.

Pastoralists are not only tasked with herd management decisions (eg herd size, species composition, transhumant migration, and pasture rotation), but also water management decisions. Without access to healthy bofedales, pastoralism in the Andes would not have endured over millennia, nor would it have become integral to Andean society, including pre-Inka, colonial, and present populations. Pastoralists directly manage water to ensure adequate forage. The alpaca, in particular, is dependent upon the forage provided by bofedales, while the llama has a broader palate for grazing dry grasses (eg Festuca spp) characteristic of the puna (Baied and Wheeler 1993). Many Andean herding practices have transformed in the postcolonial era to include nonnative domesticates, especially in the northern Central Andes (eg Cordillera Blanca, Peru), including cattle and horses, which require supplemental forage and higher labor investments, and result in greater damage to vegetation in bofedales than traditional herds.

Irrigation (*irpa*) is a fundamental aspect of *bofedal* management among pastoralists (Flores-Ochoa 1977; Palacios-Ríos 1977, 1996). Traditional practices range from the use of ephemeral canals and miniature reservoirs to the installation of intricate waterways, artificial ponds, and constructed pipelines (natural and cement). The "opening" of natural water valves (eg springs or brooks) or artificial output (eg wells), and the distribution and duration of water flow are often managed according to community-based rules and regulations. The availability of labor, infrastructure, and community-scale management of irrigation practices will vary across Andean communities and linkages with municipal, state, and national agropastoral policies and markets (Guillet and Mitchell 1991; Guillet 1992).

Communities that maintain a certain level of cohesion of pastoral identity will prioritize community-scale efforts to improve and sustain their *bofedales*. This may include annual canal cleaning (eg *faena*), pasture rotation (eg *j'akas*), shared access to community grazing plots (eg *machajes*), and seeding and transplanting *bofedal* plants. Individualized water management practices may include subtle diversions of water flow and direction, using boulders, rocks, gullies, or artificially built diversions, to evenly distribute water across a *bofedal* and to direct it to areas experiencing desiccation. Traditional knowledge of local plants allows them to be used as indicators of *bofedal* health that inform irrigation needs, including soil moisture and water table level, and signals of grazing pressure that inform herd management decisions.

Local pasture management may also include the application of fire, with diverse socioecological incentives, including clearing encroaching dry grasses, revitalizing tussock growth, communicating messages over long distances, or signaling land ownership in an area of dispute. Other herder practices can include the application of *abono* (camelid dung) as a natural fertilizer. Through pasture rotation or facilitated access to multiple grazing parcels, herding animals contribute to seed dispersal and fertilization through dispersion of droppings (Yager et al 2008).

In addition to the technical tasks of *bofedal* management, ritual performance is a vital practice associated with sociohydrological systems (Lansing 1991; van Kessel 1997; Prieto 2016). Andean ritual practices, often including glaciers and mountain springs, create bonds between social groups and the nonhuman world (Castro and Aldunate 2003), wherein relationships remain in *ayni* (reciprocity) through practices of *pagos* (payments), *cariño* (affection), and *respeto* (respect). Rituals to ensure water, *bofedal*, and animal health and abundance can include *despachos* (offerings), canal cleaning ceremonies, and pilgrimages to mountains peaks that are recognized as protective spirits (*apus*), corresponding to astronomical–cosmological cycles and the agropastoral calendar (Sallnow 1987).

Ethnographic testimonies indicate that a continuous supply of water is essential to maintain, improve, and extend a bofedal (Palacios-Ríos 1977; Yager 2015). Frequent drought conditions, or drastic reductions of water inputs, will result in bofedal degradation and even rapid loss. Though bofedales endure interannual drought conditions, dating over several millennia (eg \sim 7 ky; Hribljan et al 2015), herders attest that drought conditions are more prevalent at present day, and seasonal precipitation is noncontinuous and characterized as extreme events (eg isolated downpour events that cause greater erosion and runoff). Under current climate change conditions, herders recognize that the natural water sources (eg springs and glaciers) on which they normally depend are decreasing (Orlove et al 2008; Yager 2015; Yager et al 2019). While glacier outflow may have once created the water flow to harness and supplement bofedal replenishment, many glaciers are past peak outflow (Mark et al 2017), and accelerated loss has resulted in disconnected, patchy, and fragmented ecosystems (Seimon et al 2017). In addition to decreasing water supplies, decline in irrigation management is often due to labor shortages resulting from outmigration (for work and education) and loss of traditional knowledge across generations (Turin and Valdivia 2011). Under current climate change, bofedales are already experiencing rapid desiccation, and many are vulnerable to irrevocable loss in less than a decade. When a herder is unable to manage water flow in a bofedal, it can rapidly decay within a few years' time, especially under drought conditions. Once the bofedal is disconnected from sustaining water supplies, the restoration of these systems within a single lifetime is often not feasible, as many were constructed and maintained with water inputs, both human and natural, permitting their growth over hundreds to thousands of years.

Reframing bofedal change

While broader national and international policy is critical to mitigating the impacts of climate change on mountain water towers, local actions are also necessary to address water sustainability. Some studies seek to identify and define critical hydrological inputs to *bofedal* sustainability, yet they continue to prioritize climatological parameters in the context of projected warming over the role of local human agency. Social processes, such as migration, loss of traditional practices, and political invisibilization of pastoral communities, also lead to system outcomes of *bofedal* degradation and loss, which often reflect larger processes of environmental injustice and dispossession. When herders migrate to urban areas, impacted by economic and policy linkages, they stop managing *bofedales*, often resulting in radical consequences for mountain sociohydrological systems.

This consideration invites us to re-examine proximate (both social and natural) and underlying causes related to structural institutional, political, economic, and social processes of *bofedal* change. For example, Lima et al (2016) stated that a change in climatic variability was a crucial factor in the depopulation of the *puna* region in northern Chile. Today, many pastoralists in highland regions must manage radical changes in both climate and social systems. A transdisciplinary approach (which considers the social production of *bofedales*) both characterizes "natural" drivers of highland depopulation and considers the possibility of social processes influenced by nonenvironmental factors (eg proletarization processes, forced migration) that also lead to environmental change in the *puna* highlands.

Natural resources management policies that view bofedales as separate from human practices will overlook the imperative role of herders in conservation planning. Even within protected areas, increasing trends of land fragmentation and fencing, loss of community-based water management, and increasing privatization have led to bofedal degradation (Yager et al 2019). When institutional arrangements ignore local knowledge and practices and present them as terra nullius (land without owners), landtenure rights among pastoralists are threatened. An example of this latter situation has led to dispossession of herder's land (or land access) and violation of indigenous ancestral rights (Verzijl and Quispe 2013). Promarket managerial policies have further radicalized this effect. When applied in biased economic policy, water rights have been denied to local communities in favor of extractive users (Prieto 2015). Conservation and legal discourses that present bofedales as pristine ecosystems tend to overlook local managerial practices, and even forbid some of them (Dransart 2002; García et al 2021). They conceive herder's activities as separate from nature and picture them as an external threat rather than necessary for bofedal sustainability.

Extractive activities (eg mining and water supply to companies) are a major threat to bofedales (Castro 1997; Verzijl and Quispe 2013; Scheihing and Tröger 2017; Prieto et al 2019; Cabanillas-Trujillo and Madrid-Ibarra 2020). In particular, many extractive companies are required to evaluate the impacts of implementation on bofedales through technical studies of environmental impact assessments, which are often used to justify bofedal removal for exploration and extraction tests. These studies are conducted by contracted consulting firms that do not have advanced knowledge of mountain ecosystems, neglect linkages between bofedales and water security, and ignore the role of herders and pastoral livelihood resources.

Furthermore, rather than conducting rigorous research and environmental protocols, they respond to institutional agendas and nonarticulated macroframeworks that prioritize industrial development for export over mountain ecosystem services and local communities. In addition, the physical perturbations of mining activities (both legal and illegal, at large and small scale) affect bofedales by pumping groundwater from them, modifying the hydrological flows to systems (eg channeling water in pipes), and building infrastructure (eg roads that cut off water flow to bofedales). A paradigmatic illustration is the Chilean case of the Chuquicamata Mine (the largest open-pit copper mine in the world), which has caused irreversible destruction to several bofedales located in the San Pedro de Inacaliri river basin that had sustained indigenous communities since pre-Hispanic times (Prieto et al 2019). The brine mining boom (eg lithium, potassium, borax) is also a current threat to bofedales located near salt flats. Brine mining employs an industrial process that can be understood as a form of water mining (Garcés and Alvarez 2020; Bustos-Gallardo et al 2021). The encroachment of mining in pastoral regions often divides community alliances and drives migration to urban centers and population abandonment of the highlands. This, consequently, results in the loss of local bofedal management and fractured or discontinued community pastoral identity (Babidge et al 2019).

Another significant extractive activity is the expansion of dams to supply water or hydropower for rapidly growing urban populations, for which the infrastructure is frequently constructed on *bofedales* (eg Cordillera Real, Bolivia, and the Cordillera Vilcanota, Peru), many of which are critical habitats for endangered and threatened species (Seimon et al 2017). In Tacna (Peru), the water demand for urban consumption and agriculture has increased water extraction from *bofedales* located in the highlands (eg *Jachajawira* and Mauri river) (Carbonell 2002; Molina-Carpio et al 2012). In the arid regions of northern Chile, the expansion of industrial agriculture and mining has increased water extraction from afar, often sourced from highland *bofedal* systems (eg Glassner 1970; Bernhardson 1985; Romero et al 2017; Prieto et al 2019).

Reframing cross- and interdisciplinary research approaches has resulted in an expansion of our conceptual understanding of nature-culture relationships, including the concept of verticality, and realignment of its conceptual boundaries and practical applications (Murra 1985; Zimmerer 1999), as well as zonal classifications, including the biogeographic typology of biomes (eg puna) reframed through a social science lens as production zones (Mayer 2002). Similarly, we invite a conceptual reconsideration of bofedales—toward recognition of these unique biogeophysical mountain wetland systems as culturally produced sociohydrological systems—as an opportunity to initiate new research questions for expanding translational knowledge and to identify relevant and empowering local initiatives that recognize pastoralists as significant stakeholders in securing water tower assets.

We support transdisciplinary research that considers the proximate and underlying factors influencing land-use and land-cover change (Geist and Lambin 2002; Coppock et al 2017; Izquierdo et al 2018) and local decision-making (Gilles et al 2013; Valdivia et al 2013) to inform conservation planning. Pastoral management and highland landscapes are

rapidly changing across Andean communities, and identification of the multiscale processes and drivers impacting sociohydrological outcomes is necessary to secure sustainable water outcomes (Brelsford et al 2020). We propose reframing *bofedal* research to include the role of pastoralists as key managers of sociohydrological infrastructure in Andean water tower regions. In particular, we seek to further answer the following questions.

- What multiscalar factors (climate, sociopolitical, socioecological, and socioeconomic) and processes (including proximate and systemic) influence local pastoral decision-making regarding *bofedal* management?
- Are there shared dominant challenges of bofedal
 management across Andean water tower regions? What are
 some effective risk management strategies for bofedal
 management in relation to climate change and
 socioecological processes at multiple governance scales
 (including water extraction and market changes)?
- What translational knowledge, socioeconomic networks, technologies, and resources can be coproduced and shared across institutions, policymakers, and local stakeholders to support and extend the resilience of *bofedales* and increase the sustainability of sociohydrological systems across the Andes?

Conclusion

We posit that if the study and identification of bofedales remain limited to their biophysical dimensions, confounded as purely natural, and or left broadly defined as a mountain wetland, the outcome could result in erroneous assessments of land-use and land-cover change, which could then result in counterproductive and misinformed policies, or institutional neglect of local water rights and needs. Obfuscation of pastoralists in their roles as creators and maintainers of bofedales as key sociohydrological systems will contribute to further displacement of mountain communities, increase threats to water security on a regional scale, and cause the loss of Andean practices that have endured for millennia. Instead, we recognize that continuity of pastoral practices that support sociohydrological systems is especially critical given rapid climate change and socioeconomic processes that threaten the sustainability of mountain water towers.

Accepting the above articulated premise that pastoralists and their practices have created critical sociohydrological systems that are key components of Andean water towers, researchers and policymakers concerned with water conservation planning should reframe questions that recognize the ways in which pastoralists act as stewards of Andean water regions and identify the current and potential threats to this role. Increasing the resilience of water resources in mountain regions entails working together across disciplines and with stakeholders in order to devise actionable knowledge and practices that contribute to, rather than endanger, the sustainability of the *bofedales*.

ACKNOWLEDGMENTS

We are grateful to Drs Corinne Valdivia, Cecilia Turin, and Manuel Peralvo for valuable suggestions on the manuscript, Caroline White-Nockleby for her help with the literature review, and Arely Neisa Palabral Aguilera for the elaboration of the drawing in Figure 1. This research was funded by the Chilean National Agency

for Research and Development (Agencia Nacional de Investigación y Desarrollo de Chile, ANID), FONDECYT 1201527, FONDAP 15110006, PIA SOC180023, and PAI-MEC 80160087. Rosa Isela Meneses would also like to acknowledge support from the ANID/Scholarship Program/Doctorado Nacional (21201693).

OPEN PEER REVIEW

This article was reviewed by Cecilia Turin and Manuel Peralvo. The peer review process for all MountainAgenda articles is open. In shaping target knowledge, values are explicitly at stake. The open review process offers authors and reviewers the opportunity to engage in a discussion about these values.

REFERENCES

Babidge S, Kalazich F, Prieto M, Yager K. 2019. "That's the problem with that lake; it changes sides": Mapping extraction and ecological exhaustion in the Atacama. *Journal of Political Ecology* 26(1):738–760.

Baied C, Wheeler J. 1993. Evolution of High Andean puna ecosystems: Environment, climate, and cultural change over the last 12,000 years in the Central Andes. *Mountain Research and Development* 13(2):145–156. **Bernhardson W.** 1985. El desarrollo de recursos hidrológicos del altiplano

ariqueño y su impacto sobre la economía ganadera de la zona. *Chungara* 14:169–181.

Blaikie P. 1985. The Political Economy of Soil Erosion in Developing Countries. London, United Kingdom: Longman.

Brelsford C, Dumas M, Schlager E, Dermody BJ, Aiuvalasit M, Allen-Dumas MR, Beecher J, Bhatia U, D'Odorico P, Garcia M, et al. 2020. Developing a sustainability science approach for water systems. Ecology and Society 25(2):23. Browman D. 1989. Origins and development of Andean pastoralism: An overview of he past 6000 years. In: Clutton-Brock J, editor. The Walking Larder: Patterns of Domestication, Pastoralism, and Predation. London, United Kingdom: Unwin Hyman, pp 257–268.

Bury J, Mark BG, Carey M, Young KR, McKenzie JM, Baraer M, French A, Polk MH. 2013. New geographies of water and climate change in Peru: Coupled natural and social transformations in the Santa River watershed. Annals of the Association of American Geographers 103:363–374.

Bustos-Gallardo B, Bridge G, Prieto M. 2021. Harvesting lithium: Water, brine and the industrial dynamics of production in the Salar de Atacama. *Geoforum* 119:177–189.

Buytaert W, Cuesta-Camacho F, Tobón C. 2011. Potential impacts of climate change on the environmental services of humid tropical alpine regions. *Global Ecology and Biogeography* 20(1):19–33.

Buytaert W, Vuille M, Dewulf A, Urrutia R, Karmalkar A, Célleri R. 2010. Uncertainties in climate change projections and regional downscaling in the tropical Andes: Implications for water resources management. Hydrology and Earth System Sciences 14(7):1247–1258.

Cabanillas-Trujillo E, Madrid-Ibarra F. 2020. Impact of the mining activity of the "Humasha" exploitation project in the alto-Andean ecosystem, pampa de Coshorococha, Huayllay district, Pasco-Perú. *Biotempo* 17(1):137–162.

Canales C, Tapia Núñez ME. 1987. Producción y manejo de forrajes en los Andes del Perú. Lima, Peru: Universidad Nacionade San Cristóbal de Huamanga. Capriles JM, Tripcevich N. 2016. The Archaeology of Andean Pastoralism.

Albuquerque, NM: University of New Mexico Press.

Carbonell F. 2002. El valor de los Bofedales, humedales postergados del Perú:

Estudio de Caso en Tacna. Informe técnico de investigación. Compilación de experiencias comunitarias en el uso y conservación de la vida silvestre por pueblos indígenas, programa INFOCON de la ONG MERALVIS (Mejorando al desarrollo rural de la región a través de la conservación de la vida silvestre y el entendimiento cultural de

los pueblos). Heredia, Costa Rica: Isa Torrealba.

Castro V. 1997. Fragilidades, equilibrios y ética. Sobre patrimonios culturales y naturales. Boletín de Movimiento Agroecológico Chileno (MACH) 5(18):23–28.

Castro V. Aldunate C. 2003. Sacred mountains in the highlands of the south-

central Andes. Mountain Research and Development 23(1):73–79.

Chimner RA, Bourgeau-Chavez L, Grelik S, Hribljan J, Clarke A, Polk M, Fuentealba B. 2019. Mapping mountain peatlands and wet meadows using multi-date, multi-sensor remote sensing in the Cordillera Blanca, Peru. Wetlands 39(5):1057–1067

Chimner RA, Karberg JM. 2008. Long-term carbon accumulation in two tropical mountain peatlands, Andes Mountains, Ecuador. Mires and Peat (3)4:1–10. Cooper D, Kaczynski K, Slayback D, Yager K. 2015. Growth and organic production in peatlands dominated by Distichia muscoides, Bolivia. Arctic, Antarctic and Alpine Research 47(3):505–510.

Cooper D, Sueltenfuss J, Oyague E, Yager K, Slayback D, Cabero-Caballero EM, Mark B. 2019. Drivers of peatland water table dynamics in the central Andes, Bolivia and Peru. Hydrological Processes 33(13):1913–1925.

Cooper D, Wolf EC, Colson C, Vering W, Granda A, Meyer M. 2010. Alpine peatlands of the Andes, Cajamarca, Peru. Arctic, Antarctic, and Alpine Research 42(1):19–33.

Coppock DL, Fernández-Giménez M, Hiernaux P, Huber-Sannwald E, Schloeder C, Valdivia C, Arredondo JT, Jacobs M, Turin C, Turner M. 2017. Rangelands in developing nations: Conceptual advances and societal implications. In: Briske DD, editor. Rangeland Systems: Processes, Management and Challenges. Cham, Switzerland: Springer, pp 596–642.

Dangles O, Rabatel A, Kraemer M, Zeballos G, Soruco A, Jacobsen D, Anthelme F. 2017. Ecosystem sentinels for climate change? Evidence of wetland cover changes over the last 30 years in the tropical Andes. PLoS ONE 12(5):e0175814.

Denevan W. 2001. Cultivated Landscapes of Native Amazonia and the Andes. Oxford, United Kingdom: Oxford University Press.

Dove M. 2004. Anthropogenic grasslands in Southeast Asia: Sociology of knowledge and implications for agroforestry. *Agroforestry Systems* (61):423–435. **Dove M.** 2006. Indigenous people and environmental politics. *Annual Review of Anthropology* 35:191–208.

Dove M, Kammen D. 2015. Science, Society and the Environment: Applying Anthropology and Physics to Sustainability. Abingdon, United Kingdom: Routledge. **Dransart P.** 2002. Earth, Water, Fleece and Fabric: An Ethnography and Archaeology of Andean Camelid Herding. London, United Kingdom: Routledge.

Earle L, Warner B, Aravena R. 2003. Rapid development of an unusual peataccumulating ecosystem in the Chilean Altiplano. Quaternary Research 59(1):2– 11.

Erickson C. 2000. The Lake Titicaca basin: A pre-Columbian built landscape. *In:* Lentz D, editor. *Imperfect Balance: Landscape Transformations in the Precolumbian Americas.* New York, NY: Columbia University Press, pp 311–356.

Fairhead J, Leach M. 1996. Misreading the African Landscape. Society and Ecology in a Forest–Savanna Mosaic. Cambridge, United Kingdom: Cambridge University Press.

Figueroa-Armijos M, Valdivia C. 2017. Sustainable innovation to cope with climate change and market variability in the Bolivian Mountains. *Innovation and Development* 7:17–35.

Flannery KV, Marcus J, Reynolds RG. 1989. The Flocks of the Wamani: A Study of Llama Herders on the Puna of Ayacucho, Peru. San Diego, CA: Academic Press. Flores-Ochoa J. 1977. Pastores de Puna: Uywamichiq Punarunakuna. Lima, Peru: Instituto de Estudios Peruanos.

Gandarillas V, Jiang Y, Irvine K. 2016. Assessing the services of high mountain wetlands in tropical Andes: A case study of Caripe wetlands at Bolivian Altiplano. *Ecosystem Services* 19:51–64.

Garcés I, Alvarez G. 2020. Water mining and extractivism of the Salar de Atacama, Chile. In: Casares J, editor. Environmental Impact V. WIT Transactions on Ecology and the Environment 245. Southampton, United Kingdom: WIT Press, pp 189–199

Garcia E, Otto M. 2015. Caracterización ecohidrológica de humedales alto andinos usando imágenes de satélite multitemporales en la cabecera de cuenca del río Santa, Ancash, Perú. *Ecología Aplicada* 14:115–125.

García M, Prieto M, Kalazich F. 2021. The protection of the mountain ecosystems of the southern Central Andes: Tensions between Aymara herding practices and conservation policies. *eco.mont* 13(1):22–30.

Geist HJ, Lambin EF. 2002. Proximate causes and underlying driving forces of tropical deforestation. *Bioscience* 52(2):143–150.

Gilles J, Thomas J, Valdivia C, Yucra Sea E. 2013. Where are the laggards? Conservers of traditional knowledge in Bolivia. Rural Sociology 78(1):51–74. Glassner MI. 1970. The Rio Lauca: Dispute over an international river. Geographical Review 60(2):192–207.

Guillet D. 1992. Covering Ground: Communal Water Management and the State in the Peruvian Highlands. Ann Arbor, MI: The University of Michigan Press. Guillet D, Mitchell W, editors. 1991. Irrigation at High Altitudes: The Social Organization of Water Control Systems in the Andes. Washington, DC: American Anthropological Association.

Hartman B. 1996. Sociocultural constraints to land management decisions: The case of bofedal restoration in Bolivia. Tropical Resources Institute 15:24–28.

Hribljan JA, Cooper D, Sueltenfuss J, Wolf EC, Heckman K, Lilleskov E, Chimner R. 2015. Carbon storage and long-term rate of accumulation in high-altitude Andean peatlands of Bolivia. Mires and Peat 15(12):1–14.

Immerzeel WW, Lutz AF, Andrade M, Bahl A, Biemans H, Bolch T, Hyde S, Brumby S, Davies BJ, Elmore AC, et al. 2020. Importance and vulnerability of the world's water towers. Nature 577:364–369.

Izquierdo AE, Foguet J, Grau HR. 2015. Mapping and spatial characterization of Argentine High Andean peatbogs. *Wetlands Ecology and Management* 23:963–

Izquierdo AE, Grau HR, Navarro CJ, Casagranda E, Castilla C, Grau A. 2018. Highlands in transition: Urbanization, pastoralism, mining, tourism, and wildlife in the Argentinian Puna. Mountain Research and Development 38(4):390–400. Lane K. 2006. Through the looking glass: Re-assessing the role of agro-

Lane K. 2006. Through the looking glass: Re-assessing the role of agro-pastoralism in the north-central Andean highlands. *World Archaeology* 38(3):493–510.

Lane K. 2014. Water technology in the Andes. *In:* Selin H, editor. *Encyclopedia of the History of Science, Technology, and Medicine in Non-Western Cultures.*Dordrecht, the Netherlands: Springer, pp 1–24.

Lansing S. 1991. Priests and Programmers: Technologies of Power in the Engineered Landscape of Bali. Princeton, NJ: Princeton University Press.

Lima M, Christie DA, Santoro MC, Latorre C. 2016. Coupled socio-environmental changes triggered indigenous Aymara depopulation of the semiarid Andes of Tarapacá—Chile during the late 19th–20th centuries. PLoS ONE 11(8):e0160580. Lindsay R. 2018. Peatland classification. In: Finlayson C, Everard M, Irvine K, McInnes RJ, Middleton BA, van Dam A, Davidson NC, editors. The Wetland Book. Dordrecht, the Netherlands: Springer, pp 1515–1528.

Luebert F, Pliscoff P. 2006. Sinopsis bioclimática y vegetacional de Chile. Santiago, Chile: Editorial Universitaria.

Mark BG, French A, Baraer M, Carey M, Bury J, Young KR, Polk MH, Wigmore O, Lagos P, Crumley R, et al. 2017. Glacier loss and hydro-social risks in the Peruvian Andes. Global and Planetary Change 159:61–76.

Mathez-Stiefel SL, Peralvo M, Báez S, Rist S, Buytaert W, Cuesta F, Fadrique B, Feeley KJ, Groth AAP, Homeier J, et al. 2017. Research priorities for the conservation and sustainable governance of Andean forest landscapes. Mountain Research and Development 37(3):323–339.

Mayer E. 2002. Articulated Peasant: Household Economies in the Andes. Boulder, CO: Westview Press.

McDowell G, Huggel C, Frey H, Wang FM, Cramer K, Ricciardi V. 2019. Adaptation action and research in glaciated mountain systems: Are they enough to meet the challenge of climate change? Global Environmental Change 54:19–30. Meneses RI, Domic A, Beck S, Yager K. 2019. Bofedales Altoandinos: Oasis en la Puna. La Paz, Bolivia: The Nature Conservancy.

Meneses RI, Ortuño T, Herrera SL, Domic A, Palabral-Aguilera A, Zeballos G. 2015. Bofedales Altoandinos. In: Moya MI, Meneses RI, Sarmiento J, editors. Historia natural de un valle en los Andes. La Paz, Bolivia: Museo Nacional de Historia Natural, pp 190–205.

Mitchell WP, Guillet D, editors. 1994. Irrigation at High Altitudes: The Social Organization of Water Control Systems in the Andes. Society for Latin American Anthropology Volume 12. Washington, DC: American Anthropological Association.

Molina-Carpio J, Cruz R, Alurralde JC. 2012. Impactos transfronterizos de proyectos de trasvase: El caso de la cuenca del río Mauri. In: Proceedings of the XXV Congreso Latinoamericano de Hidráulica of the International Association for Hydro-Environment Engineering and Research (IAHR), in San José, Costa Rica, 9–12 September 2021. https://www.researchgate.net/publication/315114506_IMPACTOS_TRANSFRONTERIZOS_DE_PROYECTOS_DE_TRASVASE_EL_CASO_DE_LA_CUENCA_DEL_RIO_MAURI; accessed on 17 November 2021.

Murra J. 1985. The limits and limitations of the "Vertical Archipelago" in the Andes. In: Shozo M, Shimada I, editors. Andean Ecology and Civilization. Tokyo,

Japan: University of Tokyo Press, pp 15–20. *Orlove B, Wiegandt E, Luckman B, editors.* 2008. *Darkening Peaks: Glacier Retreat, Science and Society*. Berkeley, CA: University of California Press.

Otto M, Gibbons RE. 2017. Potential effects of projected decrease in annual rainfall on spatial distribution of high Andean wetlands in southern Peru. Wetlands 37(4):647–659.

Palacios-Ríos F. 1977. Pastizales de regadío para alpacas. *In:* Flores JA, editor. *Pastores de Puna: Uywamichiq Punarunakuna*. Lima, Peru: Instituto de Estudios Peruanos, pp 155–170.

Palacios-Ríos F. 1996. Pastizales de regadío para alpacas en la puna alta (el ejemplo de Chichillapi). In: Morlon P, editor. Comprender la agricultura campesina en los Andes Centrales, Perú y Bolivia. Lima and Cusco, Peru: Instituto Francés de Estudios Andinos, Centro de Estudios Regionales Andinos, Bartolomé de las Casas, pp 207–213.

Polk MH, Young KR, Baraer M, Mark BG, McKenzie JM, Bury J, Carey M. 2017. Exploring hydrologic connections between tropical mountain wetlands and glacier recession in Peru's Cordillera Blanca. *Applied Geography* 78:94–103.

Postigo JC, Young, K, Crews, K. 2008. Change and continuity in a pastoralist community in the High Peruvian Andes. *Human Ecology* 36:535–551. **Prieto M.** 2015. Privatizing water in the Chilean Andes: The case of Las Vegas de Chiu-Chiu. *Mountain Research and Development* 35:220–229.

Prieto M. 2016. Practicing costumbres and the decommodification of nature: The Chilean water markets and the Atacameno people. *Geoforum* 77:28–39.

Prieto M, Salazar D, Valenzuela MJ. 2019. The dispossession of the San Pedro de Inacaliri river: Political ecology, extractivism and archaeology. *The Extractive Industries and Society* 6:562–572.

Rabatel A, Francou B, Soruco A, Gomez J, Cáceres B, Ceballos JL, Basantes R, Vuille M, Sicart JE, Huggel C, et al. 2013. Current state of glaciers in the tropical Andes: A multi-century perspective on glacier evolution and climate change. The Cryosphere 7(1):81–102.

Romero H, Videla A, Gutiérrez F. 2017. Explorando conflictos entre comunidades indígenas y la industria minera en Chile: Las transformaciones socioambientales de la región de Tarapacá y el caso de Lagunillas. *Estudios Atacameños* 55:231–250.

Ruthsatz B. 1993. Flora and ecological conditions of high Andean peatlands of Chile between 18°00′ (Arica) and 40°30′ (Osorno) south latitude. *Phytocoenologia* 25:185–234.

Ruthsatz B. 2012. Vegetación y ecología de los bofedales altoandinos de Bolivia. Phytocoenología 42(3–4):133–179.

Sallnow M. 1987. Pilgrims of the Andes: Regional Cults in Cusco. Washington, DC: Smithsonian Institution Press.

Scheihing K, Tröger U. 2017. Local climate change induced by groundwater overexploitation in a high Andean arid watershed, Laguna Lagunillas basin, northern Chile. *Hydrogeology Journal* 26:705–719.

Segnini A, Carvalho JL, Bolonhezi D, Milori D, Silva W, Simões ML, Martin-Neto L. 2013. Carbon stock and humification index of organic matter affected by sugarcane straw and soil management. Scientia Agricola 70(5):321–326.

Seimon T, Seimon A, Yager K, Reider K, Delgado A, Sowell P, Tupayachi A, Konecky B, McAloose D, Halloy S. 2017. Long-term monitoring of tropical alpine habitat change, Andean anurans, and chytrid fungus in the Cordillera Vilcanota, Peru: Results from a decade of study. Ecology and Evolution 7:1527–1540.

Squeo F, Warner B, Aravena R, Espinoza D. 2006. Bofedales: High altitude peatlands of the central Andes. Revista Chilena de Historia Natural 79:245–255.

Struelens Q, Gonzales-Pomar KG, Herrera SL, Nina-Huanca G, Dangles O, Rebaudo F. 2017. Market access and community size influence pastoral management of native and exotic livestock species: A case study in communities of the Cordillera Real in Bolivia's high Andean wetlands. PloS One 12:e0189409.

Turin C, Valdivia C. 2011. Off-farm work in the Peruvian altiplano: Seasonal and geographic considerations for agricultural and development policies. *In:* Deveraux S, Sabates-Wheeler R, Longhurst R, editors. Seasonality, Rural Livelihoods and Development. London, United Kingdom: Earthscan, pp 320–335.

Urrutia R, Vuille M. 2009. Climate change projections for the tropical Andes using a regional climate model: Temperature and precipitation simulations for the end of the 21st century. Journal of Geophysical Research: Atmospheres 114:D02108. Valdivia C, Thibeault J, Gille JL, García M, Seth A. 2013. Climate trends and projections for the Andean Altiplano and strategies for adaptation. Advances in Science and Research (33):69–77.

Valdivia C, Yager K. 2018. Adapting to climate change in the Andes: Changing landscapes and livelihood strategies in the Altiplano. In: Cupples J, Prieto M, Palomino-Schalscha M, editors. The Routledge Handbook of Latin American Development. London, United Kingdom: Routledge, pp 480–499.

van Kessel J. 1997. La tecnología simbólica en la producción agropecuaria andina. In: van Kessel J, Larraín H, editors. Manos sabias para criar la vida. Quito, Ecuador: Abya-Yala, pp 33–55.

Verziji A, Quispe SG. 2013. The system nobody sees: Irrigated wetland management and alpaca herding in the Peruvian Andes. *Mountain Research and Development* 33(3):280–293.

Villagrán MC, Castro RV. 1997. Etnobotánica y manejo ganadero de las vegas, bofedales y quebradas en el Loa superior, Andes de Antofagasta, Segunda Región, Chile. Chungara 29(2):275–304.

White-Nockleby C, Prieto M, Yager K, Meneses RI. 2021. Understanding bofedales as cultural landscapes in the central Andes. Wetlands 41:102.

Yager K. 2015. Satellite imagery and community perceptions of climate change impacts and landscape change. In: Dove M, Jessica B, editors. Climate Cultures: Anthropological Perspectives on Climate Change. New Haven, CT: Yale University Press, pp 146–168.

Yager K, Resnikowski H, Halloy S. 2008. Grazing and climatic variability in Sajama National Park, Bolivia. Pirineos 163:97–109.

Yager K, Valdivia C, Slayback D, Jimenez E, Meneses RI, Palabral A, Bracho M, Romero D, Hubbard A, Pacheco P, et al. 2019. Socioecological dimensions of Andean pastoral landscape change: Bridging traditional ecological knowledge and satellite image analysis in Sajama National Park, Bolivia. Regional Environmental Change 19(5):1353–1369.

Zimmerer K. 1999. Overlapping patchworks of mountain agriculture in Peru and Bolivia: Toward a regional–global landscape model. *Human Ecology* 27(1):135–165.