

Toxicity and Affecting Factors of *Bacillus thuringiensis* var. *Israelensis* on *Chironomus kiiensis* Larvae

Authors: Cao, Chuan-Wang, Sun, Li-Li, Wen, Rong-Rong, Li, Xiao-Peng, Wu, Hong-Qu, et al.

Source: Journal of Insect Science, 12(126) : 1-8

Published By: Entomological Society of America

URL: <https://doi.org/10.1673/031.012.12601>

The BioOne Digital Library (<https://bioone.org/>) provides worldwide distribution for more than 580 journals and eBooks from BioOne's community of over 150 nonprofit societies, research institutions, and university presses in the biological, ecological, and environmental sciences. The BioOne Digital Library encompasses the flagship aggregation BioOne Complete (<https://bioone.org/subscribe>), the BioOne Complete Archive (<https://bioone.org/archive>), and the BioOne eBooks program offerings ESA eBook Collection (<https://bioone.org/esa-ebooks>) and CSIRO Publishing BioSelect Collection (<https://bioone.org/csiro-ebooks>).

Your use of this PDF, the BioOne Digital Library, and all posted and associated content indicates your acceptance of BioOne's Terms of Use, available at www.bioone.org/terms-of-use.

Usage of BioOne Digital Library content is strictly limited to personal, educational, and non-commercial use. Commercial inquiries or rights and permissions requests should be directed to the individual publisher as copyright holder.

BioOne is an innovative nonprofit that sees sustainable scholarly publishing as an inherently collaborative enterprise connecting authors, nonprofit publishers, academic institutions, research libraries, and research funders in the common goal of maximizing access to critical research.



Toxicity and affecting factors of *Bacillus thuringiensis* var. *israelensis* on *Chironomus kiiensis* larvae

Chuan-Wang Cao^{1a}, Li-Li Sun¹, Rong-Rong Wen¹, Xiao-Peng Li¹, Hong-Qu Wu², Zhi-Ying Wang^{1b*}

¹Department of Forest Protection, Northeast Forestry University, Harbin 150040, China

²Hubei Academy of Agricultural Sciences, Wuhan 430064, China

Abstract

Bacillus thuringiensis var. *israelensis* (Bti) is a suitable agent for controlling *Chironomus kiiensis*, a major pest polluting water. In this study, laboratory bioassays were used to study toxicity and affecting factors of Bti on *C. kiiensis* larvae. Tests were conducted using three commercial Bti formulations (oil miscible suspension, 1,200 ITU/mL; wettable power, 1,200 ITU/mg; technical material, 5,000 ITU/mg) of Bti. The toxicity of Bti formulations to third and fourth instar *C. kiiensis* larvae was in decreasing order of technical material, oil miscible suspension, and wettable powder, based on the 12 and 24 hour LC₅₀ values. Increasing larval densities (from 10 to 30 per bioassay cup) increased the LC₅₀ values for fourth instar *C. kiiensis* larvae. The LC₅₀ values for fourth instar larvae reared in sand substrate were higher than those from soil substrate, and autoclaved substrates significantly increased the LC₅₀ values. The technical material of Bti at 12 and 24 hours responded similarly to changes in temperature between 30° C and 15° C, but the LC₅₀ values at a range of tested temperatures showed distinct differences in time points.

Keywords: bioassay

Abbreviations: Bti, *Bacillus thuringiensis* var. *israelensis*

Correspondence: ^a chuanwangcao@yahoo.com.cn; ^b zhiyingwangnefu@yahoo.cn, *Corresponding author

Editor: Karl Gordon was editor of this paper.

Received: 25 February 2012 **Accepted:** 20 August 2012

Copyright : This is an open access paper. We use the Creative Commons Attribution 3.0 license that permits unrestricted use, provided that the paper is properly attributed.

ISSN: 1536-2442 | Vol. 12, Number 126

Cite this paper as:

Cao CW, Sun LL, Wen RR, Li XP, Wu HQ, Wang ZY. 2012. Toxicity and affecting factors of *Bacillus thuringiensis* var. *israelensis* on *Chironomus kiiensis* larvae. *Journal of Insect Science* 12:126. Available online:

<http://www.insectscience.org/12.126>

Introduction

Chironomidae (Diptera) is a species-rich family of flies, with almost 15,000 species described worldwide (Cranston 1994). As an important component of the aquatic insect community, chironomids have proven to be useful as biological indicators because of their sensitivity to chemical changes in aquatic ecosystems (Dudley and Blair 1992). In recent years, chironomid larvae have infested municipal water supply systems in more than 10 Chinese provinces (Zhang et al. 2004). Adult emergence of chironomid midges can occur at nuisance levels in areas surrounding urban and suburban aquatic habitats (Ali 1996). At a shallow wetland in northeast Florida, USA, chironomid populations have existed for several months each year at nuisance levels that interfere with human activities, necessitating control measures (Ali et al. 2008). *Chironomus kiiensis* Tokunaga has been reported to be the most prevalent chironomid species in Korea out of about 50 species recorded to date (Ree 1993). In China, *C. kiiensis* is mainly distributed in city water bodies in the southern part of the country, affecting water quality and being a potential disease vector (Lei et al. 2004). Since 2002, the biological mosquito larvicide *Bacillus thuringiensis* var. *israelensis* (Bti) has been used to control these insects. Lei et al. (2005) investigated the toxicity characteristics of Bti IPS82 on *C. kiiensis* and reported that it would be feasible to use Bti IPS82 to control *C. kiiensis* in city water sources. In this study, we describe the toxicity and affecting factors of three formulations of Bti on larvae of *C. kiiensis* to further improve biological control strategies.

Materials and Methods

Test materials

Oil miscible suspension (1,200 ITU/mL), wettable power (1,200 ITU/mg), and technical material (5,000 ITU/mg) of Bti were obtained from Hubei Kangxin Agro-industry Co., Ltd. (<http://www.btrdc.com/>).

C. kiiensis larvae

C. kiiensis egg masses were collected from water bodies in Shenzhen city, China, and were cultured in accordance with standard protocols (EPA 1993). Briefly, instead of collecting and separating eggs masses, the midges were reared in mixed-age cultures. Instar determination was achieved using length estimates. Larval midges were harvested directly from the mixed-age cultures and used in the bioassays.

Bioassay

The suspension stocks were prepared by putting 100 mg Bti into 1 L deionized water and shaking at 200 rpm for 30 minutes. Suspensions for use in bioassays were prepared by serial dilution of Bti in dechlorinated water. The acute toxicity of each compound was evaluated individually using 12 and 24 hour static toxicity tests. Ten larvae were randomly assigned to each of three replicates per treatment, giving a total of 30 larvae in each of the control and seven test treatments (0.01, 0.05, 0.1, 0.5, 1, 5, and 10 mg/L). Tests were conducted in 100-mL plastic cups containing 50 mL of test solution, which was comprised of dechlorinated water. Test vessels were not aerated and test solutions were not renewed for 24 hours. The test vessels were covered with clear plastic film to minimize evaporation. Tests were conducted at room temperature with a photoperiod of 16:8 L:D. The number of immobilized midges was recorded at 12 and

24 hours using a dissecting microscope ($\times 10$ magnification). Immobilization was defined as the cessation of all visible signs of movement or activity when viewed under the dissecting microscope. Each experiment was repeated using a different batch of midge larvae. An experiment was considered valid if mortality in the control did not exceed 10% at the end of the test and dissolved oxygen did not fall below 20% saturation. Bioassays evaluating the effects of larval age were conducted by measuring the toxicity of oil miscible suspension (1,200 ITU/mL), wettable powder (1,200 ITU/mg), and technical material (5,000 ITU/mg) of Bti on third and fourth instar *C. kiiensis* larvae. River sand and wetland soil were used to study the effects of substrate on the toxicity of Bti in this study. River sand was washed for 20 minutes in distilled water, oven dried, and passed through a 74- μ m mesh sieve onto a stainless steel tray. Topsoil was collected from Zhalong wetland, Heilongjiang, China. The basic physical and chemical characteristics of soil are shown in Table 1. Bioassay data were pooled, and a probit analysis was conducted using POLO probit analysis software. Significant differences in the resistance levels of the insects were based on non-overlapping 95% confidence intervals (Tian et al. 2011).

Results

Toxicity of Bti on fourth instar *C. kiiensis* larvae

The larvicidal activity of three Bti preparations is summarized in Table 2. The third and fourth instar *C. kiiensis* larvae were susceptible to wettable powder, oil miscible suspension, and technical material of Bti. The 12 and 24 hour LC_{50} s of the three Bti formulations showed similar changes for third and fourth instar *C. kiiensis* larvae and were in decreasing order of wettable powder, oil

miscible suspension, and technical material. The young third instar larvae were more susceptible than fourth instar larvae.

Influence of population density on the toxicity of Bti on *C. kiiensis* larvae

Table 3 shows the effect of population density on the toxicity of Bti by testing the 12 and 24 hour LC_{50} of technical material on fourth instar larvae. The 12 and 24 hour LC_{50} s of Bti with a density of 30 larvae/cup were 2.40 and 2.08 fold higher than at a density of 10 larvae/cup, respectively. The results showed that Bti activity in low densities was higher than in greater densities.

Influence of substrate on toxicity of Bti on *C. kiiensis* larvae

Table 4 depicts the influence of soil, autoclaved soil, sand, and autoclaved sand substrates on the toxicity of Bti on fourth instar *C. kiiensis* larvae. The 12 hour LC_{50} s of Bti on fourth instar *C. kiiensis* larvae residing in autoclaved soil, sand, and autoclaved sand were increased by 10.43, 9.50 and 15.39 fold compared to 10 larvae per cup in soil, respectively. Moreover, the 24 hour LC_{50} s of the fourth instar *C. kiiensis* larvae in autoclaved soil, sand, and autoclaved sand were also significantly higher than in soil substrate. However, the autoclaved substrates significantly influenced the toxicity of Bti on the *C. kiiensis* larvae. After 12 hours, the LC_{50} s for larvae from autoclaved soil and autoclaved sand were 10.43 and 1.62 fold higher than those from soil and sand, respectively.

Influence of temperature on the toxicity of Bti on fourth instar *C. kiiensis* larvae

To investigate the influence of temperature on the toxicity of Bti, five temperatures were selected to measure LC_{50} values for technical material of Bti bioassayed against fourth

instar *C. kiiensis* larvae (10 larvae per cup, sand substrate) at a range of constant temperatures. The results showed that Bti at two time points responded similarly to changes in temperature between 30° C and 15° C. Between 30° C and 27° C, the Bti at 12 hours showed a steep (3.60 fold) decline in activity, whereas the corresponding decline in activity at 24 hours was only 1.02 fold. The 12 hour LC₅₀ values at the tested temperature extremes for Bti were 3.11 mg/L (30° C) and 25.51 mg/L (15° C), while 24 hour LC₅₀ values were 0.41 mg/L (30° C) and 2.25 mg/L (15° C). The LC₅₀ values at a range of tested temperatures showed a distinct difference in time points, such as the 12 hour LC₅₀ being 23.49 fold higher than the 24 hour LC₅₀.

Discussion

Since Bti was initially described and applied to control mosquito larvae (Goldberg and Margalit 1977), Bti-based products have been used to control many mosquito (Goldberg and Margalit 1977; Margalit and Dean 1985; Amalraj et al. 2000; Fillinger et al. 2003) and chironomid species (Lei et al. 2004, 2005; Ali et al. 2008) worldwide. To date, some effective commercial Bti products have been registered for dipteran insect control (Russell and Kay 2008). For example, VectoBac WDG (3000 ITU/mg) has been recorded to be effective against *Chironomus tepperi* (Stevens et al. 2004). Formulations of Bti used at an appropriate dosage for mosquito control do not have a significant impact on most other animals or plants (Boisvert and Boisvert 2000). The chironomid species *C. kiiensis* is susceptible to three Bti formulations, with LC₅₀ values (24 hours) ranging from 0.13 to 0.37 mg/L for third instar larvae, and from 0.32 to 1.57 mg/L for fourth instar larvae. These results were similar to mosquitoes (Nayar et al. 1999) and other chironomids

(Charbonneau et al. 1994), with the susceptibility of older instar larvae to Bti declining as their age increases. Larval density can significantly affect the toxicity of Bti on mosquito larvae and chironomids (Nayar et al. 1999; Charbonneau et al. 2004). Stevens et al. (2004) also reported that increasing larval densities from 10 to 30 per bioassay cup resulted in increased LC₅₀ values for both age groups, significantly so in the case of older larvae (higher density LC₅₀ 0.80 mg/L). Our results showed a similar variation, in which the 12 and 24 hour LC₅₀s of Bti at a density of 30 larvae/cup were 2.40 and 2.08 fold higher than at a density of 10 larvae/cup, respectively. Additionally, substrate type and temperature also affected Bti efficacy. Our data demonstrated that the LC₅₀ values (fourth instar larvae) using sand substrate were higher than those in soil substrate. The characteristics of soil inhabited by chironomid larvae may therefore affect the Bti activity. Charbonneau et al. (1994) demonstrated that sediment characteristics might have an effect on the response of chironomids to Bti. Autoclaved substrates significantly increased the LC₅₀ values, thus suggesting that an increased toxicity in treated substrates might be due to the contribution of other microorganisms in original substrates. However, Stevens et al. (2004) found soil substrate reduced Bti efficacy relative to sand, and autoclaved soil substrate did not significantly affect the toxicity. These authors presumed that Bti toxins might be absorbed by clay soil particles and thus become unavailable to chironomid larvae. Previous studies observed that Cry proteins of Bt were absorbed by soil particles and became unavailable to ingestion by microbes (Venkateswerlu et al. 1992; Tapp et al. 1994; Koskella et al. 1997; Tapp et al. 1998; Pagel-Wieder et al. 2007). Therefore, further studies should investigate possibilities

of Bt toxin protein biodegradation, and how to maintain an efficient soil to target insects.

Temperature is also a major factor affecting Bti toxicity (Charbonneau et al. 1994; Nayar et al. 1999). The toxicity of VectoBac WDG and a Bti spore/crystal mixture to *C. tepperi* has been shown to display large variations in LC₅₀ values at 15° C. In this study, the data with technical materials of Bti and *C. kiiensis* larvae showed similar changes at temperatures between 30° C and 15° C. The Bti toxins can produce toxicity in the insect midgut after ingestion (Cao et al. 2010), and it is likely that an increased larval feeding activity at higher temperatures contributes substantially to the observed increases in Bti toxicity (Stevens et al. 2004). To effectively control the pestiferous chironomid larvae, it is prerequisite to optimize field conditions for maximizing the toxicity of Bti.

Acknowledgments

The work was supported by grants from the State Key Laboratory of Environmental Chemistry and Ecotoxicology, Research Center for Eco-Environmental Sciences, Chinese Academy of Sciences (KF-2008-23), Harbin Science and Technology Innovation Talents Special Foundation (2010RFQXS055), and China Postdoctoral Science Foundation (20090450942).

References

- Ali A, Lobinske RJ, Leckel RJ, Carandang N, Mazumdar A. 2008. Population survey and control of chironomidae (Diptera) in wetlands in northeast Florida, USA. *Florida Entomologist* 91(3): 446-452.
- Ali A. 1996. Pestiferous chironomidae (Diptera) and their management. In: Rosen D, Bennett FD, Capinera JL, Editors, *Pest management in the subtropics: Integrated pest management—A Florida perspective*. pp. 487-513. Intercept.
- Amalraj DD, Sahu SS, Jambulingam P, Doss PSB, Kalyanasundaram M, Das PK. 2000. Efficacy of aqueous suspension and granular formulations of *Bacillus thuringiensis* (VectoBac) against mosquito vectors. *Acta Tropica* 75: 243-246.
- Boisvert M, Boisvert J. 2000. Effects of *Bacillus thuringiensis* var. *israelensis* on target and nontarget organisms: a review of laboratory and field experiments. *Biocontrol Science and Technology* 10: 517-561.
- Cao CW, Liu GF, Wang ZY, Yan SC, Ma L, Yang CP. 2010. Response of the gypsy moth, *Lymantria dispar* to transgenic poplar, *Populus simonii* × *P. nigra*, expressing fusion protein gene of the spider insecticidal peptide and Bt-toxin C-peptide. *Journal of Insect Science* 10: 200. Available online: insectscience.org/10.200
- Charbonneau CS, Drobney RD, Rabeni CF. 1994. Effects of *Bacillus thuringiensis* var. *israelensis* on nontarget benthic organisms in a lentic habitat and factors affecting the efficacy of the larvicide. *Environmental Toxicology and Chemistry* 13: 267-279.
- Cranston PS. 1994. Introduction. In: Armitage P, Cranston PS, Pinder LCV, Editors. *The Chironomidae: Biology and Ecology of Non-Biting Midges*. pp: 1-7. Chapman and Hall.
- Dudley W, Blair WF. 1992. *Aquatic Insects*, first edition. CAB International Wallington Oxon.

- Fillinger U, Knols BG, Becker N. 2003. Efficacy and efficiency of new *Bacillus thuringiensis* var. *israelensis* and *Bacillus sphaericus* formulations against *Afrotopical anophelines* in Western Kenya. *Tropical Medicine and International Health* 8: 37-47.
- Goldberg L, Margalit J. 1977. A bacterial spore demonstrating rapid larvicidal activity against *Anopheles sergentii*, *Uranotaenia unguiculata*, *Culex univittatus*, *Aedes aegypti* and *Culex pipiens*. *Mosquito News* 37: 355-358.
- Koskella J, Stotzky G. 1997. Microbial utilization of free and clay-bound insecticidal toxins from *Bacillus thuringiensis* and their retention of insecticidal activity after incubation with microbes. *Applied and Environmental Microbiology* 63(9): 3561-3568.
- Lei P, Zhang JS, Zhou L, Pan J, Zhao WM. 2004. A study on functional characters of *Bacillus Thuringiensis* subsp. *israelensis* to *Chironomus Kiiensis* Tokunaga. *Microbiology* 31(1): 17-21.
- Lei P, Zhao WM, Yang SY, Zhang JS, Liu LJ. 2005. Impact of environmental factors on the toxicity of *Bacillus thuringiensis* var. *israelensis* IPS82 to *Chironomus kiiensis*. *Journal of the American Mosquito Control Association* 21(1): 59-63.
- Margalit J, Dean D. 1985. The story of *Bacillus thuringiensis* var. *israelensis* (Bti). *Journal of the American Mosquito Control Association* 1: 1-7.
- Nayar JK, Knight JW, Ali A, Carlson DB, O'Bryan PD. 1999. Laboratory evaluation of biotic and abiotic factors that may influence larvicidal activity of *Bacillus thuringiensis* serovar. *Israelensis* against two Florida mosquito species. *Journal of the American Mosquito Control Association* 15: 32-42.
- Pagel-Wieder S, Niemeyer J, Fischer WR, Gessler F. 2007. Effects of physical and chemical properties of soils on adsorption of the insecticidal protein (Cry1Ab) from *Bacillus thuringiensis* at Cry1Ab protein concentrations relevant for experimental field sites. *Soil Biology and Biochemistry* 39: 3034-3042.
- Pfenninger M, Nowak C, Kley C, Steinke D, Streit B. 2007. Utility of DNA taxonomy and barcoding for the inference of larval community structure in morphologically cryptic *Chironomus* (Diptera) species. *Molecular Ecology* 16: 1957-1968.
- Ree HI. 1993. Breeding places of non-biting midges (Chironomidae, Diptera) in Korea. *Korean Journal of Entomology* 23: 169-176.
- Russell TL, Kay BH. 2008. Biologically based insecticides for the control of immature Australian mosquitoes: A review. *Australian Journal of Entomology* 47: 232-242.
- Stevens MM, Akhurst RJ, Clifton MA, Hughes PA. 2004. Factors affecting the toxicity of *Bacillus thuringiensis* var. *israelensis* and *Bacillus sphaericus* to fourth instar larvae of *Chironomus tepperi* (Diptera: Chironomidae). *Journal of invertebrate pathology* 86: 104-110.
- Tapp H, Calamai L, Stotzky G. 1994. Adsorption and binding of the insecticidal proteins from *Bacillus thuringiensis* subsp. *kurstaki* and subsp. *tenebrionis* on clay-minerals. *Soil Biology and Biochemistry* 26(6): 663-679.

Tapp H, Stotzky G. 1998. Persistence of the insecticidal toxin from *Bacillus thuringiensis* subsp. *kurstaki* in soil. *Soil Biology and Biochemistry* 30(4): 471-476.

Tian L, Cao CW, He L, Li M, Zhang L, Zhang L, Liu HQ, Liu NN. 2011. Autosomal interactions and mechanisms of pyrethroid resistance in house flies, *Musca domestica*, *International Journal of Biological Science* 7(6): 902-911.

EPA. 1993. *Standard operating procedures for laboratory cultures of Chironomus tentans*. ERL-D-Sop CTI-015, Office of Research and Development, US Environmental Protection Agency.

Venkateswerlu G, Stotzky G. 1992. Binding of the protoxin and toxin proteins of *Bacillus thuringiensis* subsp. *kurstaki* on clay-minerals. *Current Microbiology* 25(4): 225-233.

Zhang RL, Wang XH, Zhou L, Zhang JS. 2004. Advances in the studies on infesting and controlling chironomids in municipal water system. *Entomological Knowledge* 41(3): 223-226.

Table 1. Physical and chemical properties of soil in Zhalong wetlands.					
Soil layer (cm)	Organic C (g/kg)	Total P (mg/kg)	Alkali-hydrolyzable N (mg/kg)	Water ratio (%)	pH
0 ~ 10	377.56 ± 17.79	21.78 ± 1.92	115.35 ± 5.04	1.49 ± 0.01	8.01 ± 0.03

Table 2. The toxicity of <i>Bacillus thuringiensis</i> var. <i>israelensis</i> to <i>Chironomus kiiensis</i> larvae.					
Larval instar	Formulation	Treatment Time (hours)	LC ₅₀ (95% CL) mg/L	Slope ± SE	χ ² (df)
Fourth	WP	12	3.18 (1.90 ~ 4.88)	2.14 ± 0.38	23.23 (19)
	WP	24	1.57 (0.79 ~ 2.56)	2.61 ± 0.53	19.31 (18)
	OF	12	3.58 (2.43 ~ 5.65)	1.52 ± 0.24	20.33 (19)
	OF	24	0.67 (0.30 ~ 1.05)	1.96 ± 0.43	18.41 (19)
	TC	12	1.30 (0.73 ~ 2.53)	1.11 ± 0.13	28.21 (19)
	TC	24	0.32 (0.16 ~ 0.53)	1.59 ± 0.27	14.25 (19)
Third	WP	12	3.25 (1.30 ~ 15.27)	0.63 ± 0.10	26.09 (19)
	WP	24	0.37 (0.14 ~ 0.68)	1.54 ± 0.27	26.85 (19)
	OF	12	2.30 (1.68 ~ 5.25)	1.65 ± 0.30	20.13 (19)
	OF	24	0.35 (0.15 ~ 0.58)	1.14 ± 0.22	17.26 (19)
	TC	12	0.92 (0.50 ~ 1.89)	0.70 ± 0.10	13.37 (19)
	TC	24	0.13 (0.03 ~ 0.27)	1.06 ± 0.19	22.51 (19)

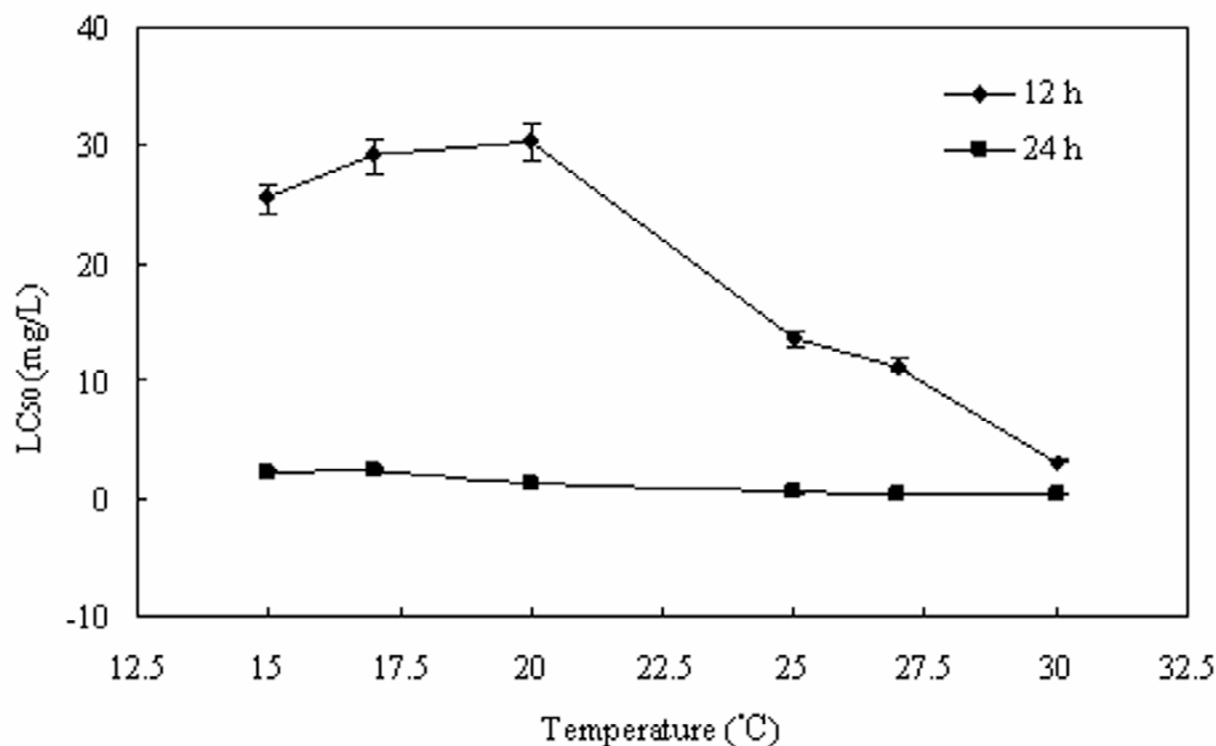
CL: Confidence Limits; WP: wettable power; OF: oil miscible suspension; TC: technical material

Table 3. Influence of population density on the toxicity of *Bacillus thuringiensis* var. *israelensis* to fourth instar *Chironomus kiiensis* larvae. The chemical used was technical material.

Larvae/cup	Treatment time (hours)	LC ₅₀ (95% CL) mg/L	Slope ± SE	χ^2 (df)
10	12	0.92 (0.50 ~ 1.89)	0.70 ± 0.10	13.37 (19)
	24	0.13 (0.03 ~ 0.27)	1.06 ± 0.19	22.51 (19)
30	12	2.21 (1.50 ~ 3.30)	0.85 ± 0.10	13.62 (19)
	24	0.27 (0.16 ~ 0.41)	1.37 ± 0.13	28.18 (19)

Table 4. Influence of substrate on the toxicity of *Bacillus thuringiensis* var. *israelensis* to fourth instar *Chironomus kiiensis* larvae at ten larvae per cup.

Treatment time (hours)	Substrate	LC ₅₀ (95% CL)	Slope ± SE	χ^2 (df)
12	Soil	9.34 (5.57 ~ 16.30)	1.03 ± 0.15	17.58 (19)
24	Soil	1.58 (0.86 ~ 2.63)	1.89 ± 0.29	28.91 (19)
12	Autoclaved soil	97.41 (51.28 ~ 347.25)	1.52 ± 0.32	15.32 (16)
24	Autoclaved soil	6.37 (2.82 ~ 24.76)	0.87 ± 0.19	15.06 (16)
12	Sand	88.70 (53.15 ~ 219.96)	1.57 ± 0.31	9.04 (19)
24	Sand	3.30 (2.13 ~ 7.10)	1.97 ± 0.40	2.67 (16)
12	Autoclaved sand	143.74 (71.48 ~ 539.42)	1.37 ± 0.25	12.59 (19)
24	Autoclaved sand	8.36 (3.45 ~ 37.12)	1.11 ± 0.20	27.94 (19)

**Figure 1.** LC₅₀ values for technical materials of *Bacillus thuringiensis* var. *israelensis* bioassayed against fourth instar *Chironomus kiiensis* larvae (10 larvae per cup, sand substrate) at a range of constant temperatures. Error bars represent 95% confidence limits. High quality images are available online.